Multi-scale Simulations of Tropospheric Chemistry in the Eastern Pacific and US West Coast during Spring 2002

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Abstract:

Regional modeling analysis for the Intercontinental Transport and Chemical Transformation experiment in 2002 (ITCT-2K2) over the Eastern Pacific and US west coast is performed using a multi-scale modeling system, including the regional tracer model CFORS, the STEM-2K3 regional chemical transport model, and an off-line coupling with the MOZART global chemical transport model. CO regional tracers calculated on-line in the CFORS model are used to identify aircraft measurement periods with Asian influences. Asian influenced air masses measured by the NOAA WP-3 aircraft in this experiment are found to have lower ΔAcetone/ΔCO, ΔMethanol/ΔCO, and ΔPropane/ΔEthyne ratios than air masses influenced by US emissions, reflecting differences in regional emission signals. The Asian air

masses in the eastern Pacific are found to usually be well-aged (> 5 days), highly diffused, and with low

NO_y levels. Chemical budget analysis is performed for two flights, and the O_3 net chemical budgets are

found to be negative (net destructive) in the places dominated by Asian influences or clear sites, and

positive in polluted American air masses. During the trans-Pacific transport, part of gaseous HNO_3 was

converted to nitrate particle, which attributed to NO_y decline. Without the aerosol consideration, the

model tends to overestimate HNO_3 background concentration along the coast region. At the measurement

site of Trinidad Head, Northern California, high concentration pollutants are usually associated with calm

wind scenarios, implying that the accumulation of local pollutants lead to the high concentration.

Seasonal variations are also discussed from April to May for this site. A high-resolution nesting

simulation with 12km horizontal resolution is used to study the WP-3 flight over Los Angeles and

surrounding areas. This nested simulation significantly improved the predictions for emitted and

secondary-generated species. The difference of photochemical behavior between the coarse (60 km) and

nesting simulations is discussed and compared with the observation.

1. Introduction:

During April and May, 2002 the Intercontinental Transport and Chemical Transformation experiment in

2002 (ITCT-2K2) studied the airmass characteristics over the eastern Pacific and the US west coast.

NOAA WP-3 aircraft and surface measurements were performed with the objective of characterizing the

Asian inflow signal and its impact on regional air quality. During this period air masses impacted by

Asian emissions are transported to the eastern Pacific by the mid-latitude prevailing westerlies. The long-

distance transport of Asian pollutants to the west coast of North America has been studied by many

researchers, such as Jaffe et al. (1999) and Kotchenruther et al. (2001). Jacob et al. (1999) discussed the

impact of Asian emissions variation on North America air quality using a global model. The strength of

the Asian influence on North America was shown to depend on the Asian emission strength and the

transport efficiency over the Pacific Ocean. Asian emissions are comprised of anthropogenic (including biofuel) sources, biomass burning, and volcanic activity. Woo et al. (2003) estimated the biomass burning emission during the spring 2001. Carmichael et al. (2003a,b) studied the features of Asian outflow over the West Pacific and used aircraft measurements and model results to evaluate the emission estimations by Streets et al. 2003. Yienger et al. (2000) described the Asian pollutant transport to North America using CO as the criteria. The trans-Pacific transport of aerosol is also an important issue. During springtime Asian dust storms become active, and dust can be transport to the eastern Pacific and North America (Uno et al., 2001; VanCuren and Cahill, 2002). However, during the ITCT-2K2 experiment the influence of Asian dust on North American was not significant.

To study the transport and chemistry of trace gases and aerosols across the northern Pacific we used a regional chemical transport model. Regional models have an advantage over global models in their ability to use finer resolutions in the analysis, but have the disadvantage of requiring lateral boundary conditions. In this study the lateral boundary conditions were established using a multi-scale modeling system with nested models. Nesting techniques can help in the analysis from two sides: introducing external influences for relatively long-lived transported species, such as CO or O_3 (Langmann et al., 2003), and by considering high-resolution emissions for short-lived species, such as NO_x or SO_2 (Tang, 2002). High-resolution emissions can more accurately resolve near-source concentrations, and better estimate photochemical budgets (Tang, 2002).

In this paper the multi-scale model is used to help analyze aircraft and surface measurements obtained during the ITCT-2K2 experiment. In the following section the details of the model are presented. In Section 3 Asian tracer information calculated by the model is used to classify the aircraft observations into those observations with large and small Asian signals, which are then subsequently used to help quantify observed characteristics of Asian air masses over the Eastern Pacific. Specific characteristics of

1 individual aircraft flights are analyzed in Section 4, where the multi-scale model system is used to study

2 the photochemistry in air masses over the eastern Pacific as aged air masses interact with emissions from

3 North America. The WP-3 flight on May 13 over Los Angeles is analyzed as an example of the effect of

4 model resolution on model predictions. The effect of aerosols on nitrate partitioning is presented in

5 Section 5. Results from a fine-scale nesting simulation of the flight around Los Angeles are presented in

6 Section 6. A mission-wide perspective of the performance of the model comprises Section 7. Analysis of

7 the Trinidad Head observations is presented in Section 8.

8

2. Model System:

11 Characterizing the Asian influence on the chemistry of the eastern Pacific is a challenge as trace species

12 are diluted during their trans-Pacific transport, which makes it difficult to distinguish them from

13 background conditions. These air masses also mix with local sources as they move in-land. To efficiently

14 consider these processes a multi-scale model system was established. This model system includes the

15 MOZART global chemical transport model (Horowitz et al., 2003), the inter-continental chemical tracer

16 model CFORS (Uno et al., 2003), and a nested regional chemical transport model, STEM-2K3 (Tang et

17 al., ACE-Asia issue). The model domains are shown in Figure 1.

18

19 CFORS is an on-line tracer model (Uno et al., 2003) coupled with the RAMS regional meteorological

20 model. In this application CFORS was driven by NCEP reanalysis for post analysis and AVN data for

21 forecasting uses. CFORS treats chemical species as tracers assuming linear consumption. For example,

22 NO_x tracer in CFORS decays with a first-order rate:

23
$$\frac{d[NO_x]}{dt} = -k_{NOx}[NO_x] \qquad (1)$$

1   where k_{NOx} is a fixed value. In CFORS, we define a conservative NO_y (total odd nitrogen species) tracer

2   that can be transported without loss, and assume that this NO_y tracer has the same emission source as

3   NO_x. Under this assumption, the NO_x age can be derived from equation (1):

4    $$T_{NOx} = k_{nox} \ln([NO_y]/[NO_x]) \qquad (2)$$

5   This equation allows for the on-line calculation of the averaged NO_x age when air masses from different

6   sources mix together. This NO_x age represents a combined result of transport time, diffusion, and NO_x

7   source intensities. With the same method we also define a VOC (volatile organic compounds) age using

8   ethane as an indicator that is related to ethane emission and decay rate. In this study, the decay rates used

9   in CFORS for CO, SO_2, NO_x and ethane were 2.22×10^{-7}, 2.78×10^{-6}, 9.26×10^{-6} and 1.25×10^{-7} s^{-1},

10   respectively.

11

12   CO is one of the primary tracers in CFORS and is used to help classify emission source types and regions.

13   In this study CO was parsed into: Asian anthropogenic, biomass burning (BB), Mexico, Canada,

14   California, Washington-Oregon, and the rest of the USA. These regional CO tracers can be used to help

15   determine the airmass properties and its mixing state. During the ITCT-2K2 field campaign forecast

16   products for these tracers were also used for flight planning. It is important to note that these CO tracers in

17   CFORS do not yield a total CO value that can be quantitatively compared to observed values because they

18   represent primary CO only. The CFORS analysis does not consider CO that arises from methane and non-

19   methane hydrocarbon oxidation. Furthermore the background levels were set to zero. In this paper,

20   CFORS tracer model for ITCT-2K2 used a domain with 100×42 grids with a 200 km horizontal resolution

21   (Figure 1), which covers East Asia, the northern Pacific Ocean, and most North America.

22

23   Comprehensive chemistry and transport interactions are calculated by the STEM-2K3 regional chemical

24   transport model, which is a further development of the STEM-2K1 model (Tang et al., 2003a; Carmichael

25   et al., 2003a) that includes the SAPRC-99 gaseous mechanism (Cater, 2000) and an explicit photolysis-

5

1   rate solver (the on-line TUV (NCAR Tropospheric Ultraviolet-Visible radiation model, Madronich and

2   Flocke, 1999)). The main improvement of STEM-2K3 over STEM-2K1 is that the former also includes an

3   aerosol thermodynamics module, SCAPE II (Simulating Composition of Atmospheric Particles at

4   Equilibrium) (Kim et al, 1993a, b; Kim and Seinfeld, 1995), for calculating gas-particle equilibrium

5   concentrations among inorganic aerosol ions and their gaseous precursors. Tang et al. (ACE-Asia issue)

6   described the framework of STEM-2K3 and its performance during the TRACE-P (TRAnsport and

7   Chemical Evolution over the Pacific) (Jacob et al., 2003) and ACE-Asia (Aerosol Characterization

8   Experiment in Asian Pacific Region) experiments. In this paper the analysis includes inorganic aerosols in

9   4 size bins (in diameter): 0.1µm-0.3µm, 0.3µm-1.0µm, 1.0µm-2.5µm, and 2.5µm-10µm, (referred to as

10   bins 1 to 4, respectively). Daily TOMS (Total Ozone Mapping Spectrometer) data was used to calculate

11   the model's overtop ozone column needed in the photolysis calculations using the TUV module.

12

13   The primary domain for STEM-2K3 covers the U.S. west coast and eastern Pacific with a 60km

14   horizontal resolution (Figure 1). Another RAMS model run in this domain was used to drive STEM-2K3.

15   To better reflect the influence of more local sources, such as during the flight over the Los Angeles basin,

16   a higher resolution is necessary. Here a nested domain with a nesting ratio of 1:5 (Figure 1) was used. The

17   nested STEM-2K3 was driven by meteorology interpolated from the RAMS 60km predication. The

18   change in resolution affects the emission intensity. For example, Figure 1 shows that CO emission rate in

19   Los Angeles ranges from 150 to 300 mole/hour/km^2 represented in the domain of 200 km resolution, is

20   over 1500 mole/hour/km^2 in the 60 km domain, and greater than 3000 mole/hour/km^2 (peak value) with

21   12 km resolution. Tang (2002) discussed how model resolution affected O_3 and NO_x predications, and

22   photochemical correlations through for the TRACE-P and ACE-Asia experiments.

23

24   **2.1 Lateral Boundary Conditions**

1   The MOZART global model was used to provide lateral boundary conditions for the STEM-2K3 primary

2   60km domain (Figure 1) applications. The lateral boundary and top boundary conditions of STEM-2K3

3   for all species except CO were interpolated to the primary 60 km domain from the MOZART model

4   results produced using NCEP winds. Due to the uncertainties in the CO emission inventories the global

5   model tends to quantitatively underestimate CO in the mid latitudes and altitudes. The ITCT-2K2 aircraft

6   observations showed that CO had a stable background concentration over the Eastern Pacific, and CO

7   contributions from various sources were mostly represented by enhancements above background. In this

8   paper, we set the boundary condition of STEM-2K3 for CO to the background CO (shown in Table 1)

9   plus a perturbation calculated as the total tracer CO from the CFORS model. The boundary values above

10   the flight altitude of the WP-3 were obtained using the observations by the NASA DC-8 aircraft during

11   the TRACE-P experiment. Table 1 also shows the model vertical layers used in CFORS and STEM-2K3,

12   defined in the mid-point of the RAMS sigma-z layers (Pielke et al., 1992). The nested 12km domain has

13   the same vertical layers as the 60 km primary domain, and the lateral boundary conditions were also

14   interpolated from the 60 km domain.

15

16   **2.2 Emissions**

17   Emission data used in this study came from various sources. The basic strategy to develop the assembled

18   emissions inventory was to use global scale emissions inventories as background/lower quality assurance

19   information, and to supplement these with more comprehensive/higher quality regional scale emissions

20   data for Asia and North America regions. The Asian anthropogenic emissions were based on the estimate

21   of Streets et al. (2003a; 2003b). Biomass burning emissions for Southeast Asia were based on April-

22   averaged Asian BB emissions for the base year of 2001 (Woo et al., 2003). Emissions in the United State

23   and Canada were based on the USEPA 1996 inventory. Mexican emissions came from Kuhns (2001). The

24   ship emissions for CO, SO_2 and NO_x were based on the inventory of Corbett et al. (1999), and aviation

25   emissions were taken from EDGAR (Olivier et al., 1996). Lightning NO_x emissions were diagnosed from

1 the meteorological model according to deep convective intensities (Pickering et al., 1998). Emissions for

2 all other regions in the 200 km domain came from GEIA inventory (Global Emissions Inventory

3 Activities, http://geiacenter.org/). Biogenic emissions for the regions other than the United States come

4 from GEIA (Guenther et al., 1995). MOZART global model has it own emission (Horowitz et al., 2003),

5 which is mainly based on the EDGAR inventory (Olivier et al., 1996).

6

7 For simulating pollutant re-circulation from the western U.S. into the Pacific Ocean and propagation from

8 the West Coast inland, we used US EPA emission databases and more detailed allocation procedures. For

9 example, the emission data for the nested 12km domain (Figure 1) requires a resolution higher than

10 county scale. The original county-scale emissions were redistributed to the finer resolution grid system

11 using high-resolution population data to give "within county" spatial variability. Further detail into the

12 methodology for this approach used to generate high-resolution emissions for nested simulations can be

13 found in the references elsewhere (Woo et al., 2003; Tang, 2002).

14

3. Air-mass Characteristics

16 The NOAA WP-3 aircraft performed 13 flights, including transit flights to and from the base in Monterey,

17 California (Figure 1). One objective of these flights was to characterize Asian air masses over the eastern

18 Pacific and how they are modified as they move over the American continent and mix with local sources.

19 Identification of air masses impacted by Asian emissions is difficult due to the long transport times and

20 the variety of sources impacting the Pacific Basin. Techniques for identifying Asian emission signals that

21 rely on observation-based filters are discussed in Nowak et al. and de Gouw et al. (this issue).

22 Alternatively the model can be used to identify those air masses expected to contain Asian signals. We

23 used an Asian air-mass ratio, defined as the Asian anthropogenic CO tracer concentration divided by total

24 anthropogenic tracer CO concentration in this domain, to identify air masses impacted by Asian sources.

25 This metric was calculated for each WP-3 3-minute flight segment, and the aircraft data was then sorted

1   using the model calculated Asian ratios. Figure 2 shows correlations of observed species for air masses

2   with Asian ratios < 20% and those with values > 80% using the 3-minute merged data set for all ITCT

3   WP-3 flights. The data points with Asian ratio < 20% and > 80% account for about 23% and 53%

4   respectively in all WP-3 flights. Figure 2 also shows the airmass correlations of the observations taken on-

5   board the NASA DC-8 and P-3B aircrafts over the Western Pacific during the TRACE-P experiment,

6   March of 2001. Figures 2a and 2b show that Asian air masses have lower $\Delta NO_y/\Delta CO$ values in both the

7   Western Pacific and Eastern Pacific than air masses dominated by American emissions. Tang et al.

8   (2003b) found that BB plumes from Southeast Asia have low $\Delta NO_y/\Delta CO$ ratios (about 0.005 ppbv/ppbv).

9   Asian air masses impacted by biofuel sources exhibit similar low ratios. The portion of emissions from

10 gasoline in Asia is generally smaller than that in the United States (Streets et al., 2003). For example,

11 coal is the main fossil fuel used in China, and coal combustion also emits a lower $\Delta NO_y/\Delta CO$ than

12 gasoline combustion. All these factors contribute to Asian air masses having a ~10 times lower

13 $\Delta NO_y/\Delta CO$ ratio than American air masses. After long-distance transport the ratio in the Asian air masses

14 decreases to ~0.0035 ppbv/ppbv (Figure 2a) due to NO_y's gas-particle conversion, and wet and dry

15 depositions. This is discussed further in Section 5.

16

17 The air masses dominated by American sources show higher correlation coefficients than Asian air

18 masses for CO versus acetone and methanol (Figures 2c-2f). These results imply that CO, acetone and

19 methanol in the United States come from similar emission sources. Air masses with Asian ratios < 20%

20 have higher $\Delta Acetone/\Delta CO$ and $\Delta Methanol/\Delta CO$ ratios, reflecting the large petroleum refining and usage

21 in the western USA.

22

23 The lifetimes of propane and ethyne range from several days to several weeks. However, they have

24 similar reaction rate with OH, the main reaction for their consumption. Figures 2g and 2h show that the

25 $\Delta Propane/\Delta Ethyne$ ratio associated with Asian sources (1.2 ppbv/ppbv) is 70% lower than in American

1   air masses, but still higher than the TRACE-P measurement since Asian air masses in the eastern Pacific

2   are mixed with North American sources.

3

4   The values presented above based on a model-based screening for Asian plumes can be compared to

5   values calculated based on observation-based plume identification. An observation-based Asian plumes

6   value of $\Delta NO_y/\Delta CO$ for the May 5 flight for altitudes > 5km of 0.003 ppbv/ppbv was reported by de

7   Gouw et al. (this issue), while values of 0.0035-0.0049 ppbv/ppbv for all identified Asian plumes were

8   calculated by Nowak et al. (this issue). In terms of $\Delta Propane/\Delta ethyne$, observed values ranged from 1.5

9   to 1.1 ppbv/ppbv for the May 5 and May 17 flights, respectively, which represent the two most heavily

10 influenced Asian events. These results indicate that the model-based approach provides a consistent and

11 4-dimensional contextual method to help identify Asian air masses.

12

13 The observations also provide indicators of ozone production. NO_z (NO_y-NO_x) represents the oxidized

14 products of NO_x, including peroxyacetyl nitrate (PAN), HNO_3, HNO_2 et al. The ratio $\Delta O_3/\Delta NO_z$

15 represents the upper limit of the ozone production efficiency (OPE) per unit NO_x (Trainer et al., 1993).

16 Figure 2i shows this ratio for the air masses with Asian ratios < 20% and > 80%, and O_3 and NO_z have

17 relatively strong correlations in both sets of data. The $\Delta O_3/\Delta NO_z$ ratio in Asian air masses over the

18 Eastern Pacific is higher than that in American air masses and in air masses over the Western Pacific

19 (Figure 2j). The very high $\Delta O_3/\Delta NO_z$ of Asian air masses over the Eastern Pacific (Figure 2i) is not solely

20 due to the accumulation of NO_x photochemically generating O_3. The NO_z conversion to nitrate aerosol

21 and depositions also results in an increase in this ratio. This was shown to be important in the Asian

22 outflow during ACE-Asia, and where most of the nitrate was concentrated in the super-micron particles,

23 and thus were more rapidly removed from the air via deposition processes (Tang et al., 2003).

24

1   Classification of the air mass age is an important element of analysis. Air mass age can be estimated from

2   trajectory analysis (Cooper et al., this issue). This can also be done using observed chemical ratios as

3   discussed by de Gouw et al. (this issue). The multi-scale model also includes indicators of chemical age.

4   Figure 3 shows the observed Propane/Ethane and NO_x/NO_y ratios plotted against predicted CFORS VOC

5   and NO_x ages, respectively. The aircraft measurements indicate that fresh air masses from the United

6   States have Propane/Ethane ratios > 1, but air masses with ages > 50 hours have ratios < 0.4. The main

7   consumption of propane and ethane in the atmosphere is through reactions with OH, and the reaction

8   constant for propane is about four times bigger than that for ethane. Figure 3a shows that the

9   Propane/Ethane ratio declines very fast within the first 50 hours. After 50 hours this ratio takes values

10   between 0.02-0.3. The relation between observed NO_x/NO_y and predicted NO_x age (Figure 3b) is similar

11   to that between Propane/Ethane and VOC age. The best-fit lines in Figure 3 show the lower limits to be

12   0.1 for Propane/Ethane, and 0.08 for NO_x/NO_y.

13

14   The points in Figures 3a and b are color-coded using CFORS VOC age. The results for the NO_x age show

15   that the two age metrics produce consistent results. The fraction of NO_x in NO_y shows more variability at

16   a given age than the Propane/Ethane ratio, which reflects the complexity of the NO_x chemistry relative to

17   that for propane and ethane. In fresh polluted air masses NO_x usually is converted to NO_z through

18   photochemical reactions, and O_3 is one of the by-products during this conversion. However, some

19   reactions, such as the photolysis of PAN, recycle NO_z back to NO_x. The ITCT measurements indicate that

20   some NO_y enhancements at high altitudes were associated with stratospheric intrusions, implying that

21   stratospheric contribution to NO_y or NO_x via N_2O dissociation (Strahan, 1999) could be significant for

22   these flights.

23

24   **4. Case Studies**

25

The multi-scale modeling system was used in forecast-mode for flight planning, and in post analysis

mode to help interpret the observations. The air mass age discussion in the previous section indicated that

the air masses from Asia took on the order of one week to reach the Eastern Pacific.   During this

transport journey dynamic mixing and dispersion processes act to decrease the air mass concentrations.

This fact, coupled with the complexity of the interactions between aged air masses and fresh emissions

from the western US, make it difficult to interpret the observations. Here we show how the various model

products can aid in the analysis and how the observations help to evaluate the model performance. These

aspects are explored through the analysis of individual flights. The flights discussed above focus on

typical flights that involve a mixture of local and Asian influences.

4.1 WP-3 Flight on April 25

On April 25, the NOAA WP-3 aircraft performed its second flight along the California coast (Figure 4),

and encountered aged Asian air masses mixed with American pollutants. In the model's 1 km layer over

most of the domain, with the exception of the continental US, the anthropogenic CO Asian ratio is higher

than 0.8 (80%) (Figure 4a). Figure 4b shows that Asian sources have a stronger impact on CO over North

America at higher altitudes. Figures 4c and 4d show simulated O_3 over the eastern Pacific and the WP-3

flight path, which was located near a low-pressure system. At low altitudes (Figure 4c) O_3 was formed

over the western US and transported out over the ocean, as shown by the band with $O_3 > 55$ ppbv that

extends southwestward from northern California. The oceanic area with $O_3 > 45$ ppbv west of Mexico is

associated with an aged polluted airmass comprised of a mixture of Asian and North American air (as

shown by the Asian ratio <0.6 as shown in Figure 4a). In the higher altitudes ozone concentrations were

mainly maintained by downward transport of stratospheric ozone (Figures 4d and e).

Figure 4e shows the cross–section of O_3 concentration and chemical net budget along latitude 35°N, a

section that approximates the path of the WP-3 flight 2. Simulated O_3 showed a net loss over the ocean

1   areas and above 3 km over land, and a net production below ~3 km over the polluted regions near Los

2   Angles. In the low altitudes over the ocean, O_3 had the highest loss rates and lowest concentrations

3   because the high water vapor conditions within the marine boundary layer (MBL) benefited the O^1D +

4     $H_2O \rightarrow 2OH$ reaction, which reduced the chance of O^1D converting back to O_3 via $O^1D + Air \rightarrow O^3P$ and

5     $O^3P + O2 \rightarrow O_3$. Due to the high OH production in the MBL the CO net chemical budget was ~ -0.3

6   ppbv/hr (Figure 4f). The CO chemical budget became positive only over the polluted regions where

7   hydrocarbon oxidization was important.

8

9   Figure 5 compares measurements of WP-3 flight 2 with simulated values extracted along the flight path.

10   The tracer-CO results are plotted in Figure 5a. The CO Asian ratios varied by height, with values > 65%

11   and <40% for the flight altitudes higher than 3 km and lower than 2km, respectively. Asian air masses are

12   usually transported across the Pacific in the mid to upper troposphere, and brought to lower altitudes

13   through convective exchange and subsidence. At low altitudes over inland locations, such as the flight

14   path around 18:55 GMT, the CO Asian ratio fell below 10%. For short-lived species, like NO_x and SO_2,

15   the Asian influence is small due to removal and conversion processes during the long-distance transport.

16   Figure 5b shows that the STEM simulation was able to capture the main observed ozone features. The

17   calculated ozone net photochemical budget is also presented and the net budget reached a maximum

18   positive value around 18:55GMT, when the aircraft flew inland at low altitudes. At other flight locations

19   the O_3 chemical budget was negative.

20

21   Figure 5c shows that modeled CO exhibits much less variability than that observed, and underestimated

22   the observed CO by more than 10 ppb during 19:00-21:00 GMT. The calculated CO chemical budget was

23   positive when the pollutant loading was high, reflecting CO produced from non-methane hydrocarbon

24   (NMHC) oxidation. For other locations the CO budget was always negative and CO consumption via

25   reaction with OH dominated. The chemical net budget for PAN is similar to that of CO. PAN was

1   chemical formed over polluted areas, while photolytic loss was the main reason for the negative budgets

2   at clean locations. Figure 5 also shows that the STEM model accurately predicted J-values (photolysis

3   rates) for $J[O_3 \rightarrow O_2 + O^1D]$ and $J[NO_2 \rightarrow NO + O^3P]$ using the on-line TUV. The agreement between the

4   simulation and measurement for $J[O^1D]$ is better than that for $J[NO_2]$, since $J[NO_2]$ is sensitive to long-

5   wavelength radiation, which can be affected by fractional clouds (a quantity that is difficult to model).

6   This flight encountered some fractional clouds. Figure 5f shows that several of the peak values for $J[NO_2]$

7   above 4km were not predicted by the model and these could be associated with fractional clouds. Similar

8   results were found for J-values over Western Pacific during TRACE-P as discussed by Tang et al.
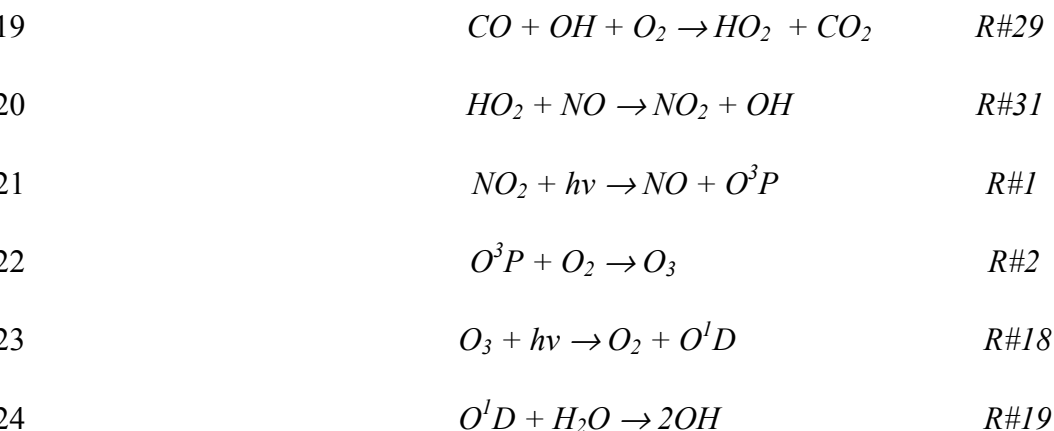
9   (2003a).

10

11   To explore the photochemical system during this flight three points along the flight path were further

12   analyzed. The WP-3 aircraft observations at 19:00, 19:03 and 21:00 GMT were selected. The points at

13   19:00 and 19:03 GMT were when the aircraft made a vertical ascent downwind of Los Angles. The model

14   simulation shows disparate air massses at the beginning and end of this profile, with the predicted CO

15   Asian ratio of 10% at 19:00GMT and 64% at 19:03GMT. When air masses from the Pacific Ocean moved

16   on-shore during the daytime an internal boundary layer (IBL) formed (Garratt, 1990) which in the model

17   extended to ~2 km. Within the boundary layer the air was mixed with local pollutants transported from

18   Los Angeles and surrounding area. Above this boundary layer, the air preserved its aged Asian features.

19   The point at 21:00GMT was at the west end of this flight at low altitude over the Eastern Pacific.

20

21   Figure 6 shows the simulated chemical production and loss terms for O_3, CO, OH, HO_2 and HCHO at

22   these three points. The O_3 budget components (Figure 6a) show that O_3 had a positive budget and the

23   highest photochemical generation through $O^3P + O_2 \rightarrow O_3$ (R#2 in Table 2) at 19:00GMT in the polluted

24   airmass. O_3 also had a lower photolysis rate (R#17 and R#18 in Table 2) at 19:00GMT than at 19:03GMT

25   due to the altitude dependency of the actinic flux. At 19:03GMT the O_3 net budget was near zero. When

1   the WP-3 aircraft flew at low altitude over the clean oceanic area at 21:00GMT the O_3 budget was

2   negative because of the rapid consumption of O^1D in the water rich MBL via $O^1D + H_2O \rightarrow 2OH$ (R#19

3   in Table 2), as was also discussed for Figure 4e.

4

5   Under these conditions there is higher OH production (Figure 6c) and CO net loss (Figure 6b) at

6   21:00GMT. Figure 4b shows that the polluted airmass encountered at 19:00GMT had the highest CO

7   consumption through R#29: $CO + HO + O_2 \rightarrow HO_2 + CO_2$ due to its high OH and CO concentrations.

8   However, this polluted point also had the highest CO production among these three points due to

9   oxidation and photolysis of aldehydes. The simulation indicates that most of the aldehydes were

10   intermediate products of hydrocarbon oxidization at 19:00GMT, and that the hydrocarbon oxidation offset

11   part of CO loss, resulting in a net positive CO budget in the polluted area (Figure 4f). The two clean

12   locations at 19:03GMT and 21:00GMT had similar low concentrations for most hydrocarbons and

13   aldehydes, but the OH production in the MBL at 21:00GMT was high, resulting in the highest CO net loss

14   among these three points.

15

16   Figure 6c shows that the polluted point of 19:00GMT had the highest OH concentration, mainly due to

17   the additional OH production through R#31: $NO + HO_2 \rightarrow OH + NO_2$. In polluted areas the following

18   reactions are important:

19   $$CO + OH + O_2 \rightarrow HO_2 + CO_2 \qquad R\#29$$

20   $$HO_2 + NO \rightarrow NO_2 + OH \qquad R\#31$$

21   $$NO_2 + hv \rightarrow NO + O^3P \qquad R\#1$$

22   $$O^3P + O_2 \rightarrow O_3 \qquad R\#2$$

23   $$O_3 + hv \rightarrow O_2 + O^1D \qquad R\#18$$

24   $$O^1D + H_2O \rightarrow 2OH \qquad R\#19$$

1   This series of reactions is constrained by NO_x and NMHC concentrations. Under optimal NO_x and

2   NMHC conditions these reactions effectively produce OH. The point at 19:00GMT reflects this situation,

3   and high OH concentrations were observed downwind of Los Angeles, which will be discussed later.

4   Under low NO_x conditions these reactions can not be performed completely, and Figure 6c shows that the

5   reaction rates of R#31 at 19:03GMT and 21:00GMT were much lower than that at 19:00GMT. The

6   reaction rate of R#19 was highest at 21:00GMT due to its rich water vapor, which resulted in elevated OH

7   concentrations, though still lower than that at the polluted point. At the polluted point OH increased by

8   converting HO_2 to OH. Here HO_2 concentrations were maintained (Figure 6d) by HO_2 production via

9   reactions R#46, R#51, R#123 and R#125 that converted RO_2 and HCHO to HO_2. RO_2 and HCHO were

10   produced as intermediate products of hydrocarbon oxidization. Figure 6d shows that the MBL site had the

11   highest HO_2 concentration, due to high OH (for R#29) and low NO_x (for R#31) concentrations. Figures 6c

12   and 6d illustrate the impact of fast conversion between OH and HO_2, and their connection to external

13   sources. R#19 was the main external source for HO_x at the MBL, and RO_2 and aldehydes were the main

14   precursors of HO_x at the polluted site.

15

16   Figure 6e shows that HCHO had negative budgets at all these sites, and the MBL site had the highest

17   HCHO net loss due to its high OH. The polluted point at 19:00GMT had the highest HCHO consumption,

18   but it also had the highest HCHO production from hydrocarbon oxidization. At the clean points HCHO

19   mainly came from methane oxidization $CH_4 + OH + O_2 \rightarrow CH_3O_2 + H_2O$ and

20   $CH_3O_2 + NO + O_2 \rightarrow NO_2 + HCHO + HO_2$ (R#46 in Table 2), which was constrained by NO_x

21   availability. Other production pathways, such as $CH_3O_2 + CH_3O_2 \rightarrow CH_3OH + HCHO + O_2$, were

22   usually small. The production of HCHO was not sufficient to offset HCHO consumption except in heavy

23   polluted areas. At the 19:00GMT point the pollutants were already diluted, and the HCHO net chemical

24   budget was negative.

25

4.2 WP-3 Flight on May 15

The 11th NOAA WP-3 flight flew west to 129°W over the Eastern Pacific on May 15 (Figure 7). This

flight encountered aged Asian and American air masses. Figure 7a shows that CO Asian ratio was high

over most of the Eastern Pacific in the model's 1 km layer. Some fresh air masses with Asian CO > 15

ppbv approached this region, but did not arrive at the WP-3 flight path. Along the North American

coastline CO Asian ratios showed a strong gradient. At 5.4 km (Figure 7b) the CO Asian ratio was higher

than 0.9 throughout the domain north of 30°N, which reflects the high transport efficiency in mid to high

latitudes and high altitudes. The influence of American sources was concentrated to inland areas at low

altitudes. The WP-3 flight region was ahead of a cold front, and north winds dominated (Figures 7c, 7d,

and 7e). The pre-frontal ascending winds transported low O_3 concentrations from the MBL to higher

altitudes, and the post-frontal descending air caused a high O_3 zone around 155°W, 44°N, which was

shown in all the layers (Figures 7c, 7d and 7e). It should be noted that the relatively fresh Asian air

masses (northwest corner of this domain) were also transported mainly from the high altitudes due to the

high transport efficiency.

Figures 7f, 7g and 7h show the O_3 chemical net budget in the 1 km, 2.8 km and 5.4 km layers. In the l km

layer O_3 net production was high over polluted inland areas and in the fresh Asian air masses (Figure 7f).

The distribution of O_3 concentrations (Figure 7c) shows high concentrations in regions with positive net

budgets, indicating that the O_3 chemical budget was the dominant factor affecting O_3 concentrations in the

low altitudes. In the 2.8 km layer transport effects were important as shown by the fact that the O_3

chemical net budget was negative (< -0.1 ppbv/hr) over most of California, but O_3 concentrations were

still higher than background (> 65 ppbv) over most Southern California. In the 5.4km layer transport

effects dominated. Figure 7h shows that the O_3 net budget was negative over most of this domain, but

very high O_3 concentrations (> 100 ppbv) existed in some areas due to the stratospheric influence. Under

1   these conditions elevated O_3 levels can correspond to areas with strong O_3 loss caused by photolysis, such

2   as the high-O_3 zone around 112°W, 48°N.

3

4   The CO net budget distribution is similar to O_3 in the 5.4km layer, as O_3 loss is mainly through its

5   photolysis process, and the O^1D produced forms OH (via R#19 in Table 2) which consumes CO. Under

6   these conditions O_3 loss is usually correlated with CO loss. Figures 7i, 7j, and 7k show that the CO net

7   budget is positive only over heavily polluted areas and their downwind sites rich in hydrocarbons. At high

8   altitudes the CO budget was always negative.

9

10   The WP-3 flight 11 flew a triangle flight path over the Eastern Pacific (Figure 7f), with its west end

11   longitude reaching about 129°W. In contrast to the WP-3 flight 2, this flight had some low-altitude points

12   with possible strong Asian signals (cf, 20:50GMT, Figure 8a). At higher altitudes (> 3km) the CO Asian

13   ratio was greater than 0.8. Figure 8a shows that the CO Asian ratio was lower than 30% only in the low-

14   altitude near-coast flight segments. The STEM simulated CO agrees well with the observations (Figure

15   8b), and the simulated CO budget shows that net CO production occurred only in the departing and

16   arriving segments, near the polluted San Francisco bay area.

17

18   Figure 8c shows the O_3 concentration comparison and simulated O_3 net chemical budget. The calculated

19   ozone captured many of the large-scale features (including the altitude dependency), but was not able to

20   resolve the finer scale features observed. The simulated O_3 net chemical budget for this flight is similar to

21   that of flight 2. The model accurately captured ethyne values as shown in Figure 8d. Since ethyne has no

22   photochemical sources its net chemical budget is always negative, and mainly determined by local OH

23   concentrations, even over polluted area. The variation of ethyne chemical budget is very similar to that of

24   CO over ocean since both are determined by the OH concentration. The OH concentration was very

25   sensitive to photolysis rate, and the model reasonably simulated the observed J-value behavior (Figure 8e,

1   8f). Most of the areas that the WP-3 flight 11 flew over were under clear-sky conditions, and the

2   photolysis rates were mainly affected by altitude and sunlight zenith angle. Figure 8g shows the

3   comparison for PAN concentrations and the simulated PAN chemical budget, which are similar to those

4   for O_3 (Figure 8c). The PAN formation is mainly through photochemical reaction between NO_x and

5   NMHCs, and its loss is mainly caused by its photolytic destruction.

6

7   The WP-3 flights 2 and 11 represent the situation typically encountered by most flights over the Eastern

8   Pacific and the near-coast regions. In these flights Asian air masses impacted by Asian sources dominated

9   at altitudes above 4 km, but had little effect over the inland low-altitude areas. Since aged air masses

10   usually were low in NO_x the O_3 photochemical production was low and the net chemical budgets were

11   negative in the Asian air masses arriving over North America.

12

5. Nitrate Partition

14   Tang et al. (ACE-Asia issue) discussed nitrate partitioning between the gas and aerosol phases for the

15   ACE-Asia and TRACE-P experiments, and found that the regional model without aerosol considerations

16   tended to overestimate nitrate acid (HNO_3). Similar results (Neuman et al., 2003) were found for the

17   ITCT-2K2 flights along the coast where regional pollutants interacted with sea salt, resulting in some of

18   the nitric acid partitioning into the sea salt particles. Figure 9 shows the simulated HNO_3 mixing ratios

19   with and without aerosol uptake compared to the measurements for WP-3 flights 4, 5, 8 and 9. The

20   simulation with aerosol uptake had lower HNO_3 concentrations than the simulation without aerosol for

21   most segments of these flights. Due to the relatively coarse horizontal resolution, the simulation with

22   aerosol underestimated most of the HNO_3 peak values. The most significant differences between these

23   two simulations appear for the predictions for background HNO_3 levels. The simulation without aerosol

24   uptake systematically overestimated the low background concentrations, especially for WP-3 flight 5, and

25   it overestimated the HNO_3 peak value for WP-3 flight 4. Under cation-limited conditions the simulation

1   with aerosol uptake can yield higher HNO_3 values (Figure 9b) due to competition effects involving

2   sulfates. However, over most clear oceanic sites, aerosol uptake tended to reduce gas phase HNO_3. This

3   partitioning is an important factor causing the very low NO_y ratio in the long-distance transported Asian

4   air masses (Figures 2a and 2b).

5

6. Nested Simulation for the Los Angeles plume

7   The WP-3 flight 10 performed an urban plume study over Los Angeles (LA) and surrounding areas. For

8   this flight the contribution of Asian sources were small. The primary domain over the Eastern Pacific

9   (Figure 1) with its 60 km horizontal resolution was not able to capture the fine structures caused by local

10   emissions. Tang (2002) tested the multi-scale simulation using nested domains in Nashville, Tennessee,

11   and found that nesting predictions could better reflect power plant and urban plume structures. When the

12   10th WP-3 flight flew over LA and its downwind areas it encountered strong local plumes with steep

13   concentration gradients, implying that local emissions played the dominant role on species concentrations.

14   To simulate this scenario, a domain with a 12 km horizontal nested grid within the primary Eastern

15   Pacific domain (Figure 1) was used. Simulated CO and O_3 concentrations in the coarse and fine domains

16   are shown in Figure 10. The coarse and fine simulations produced qualitatively similar CO and O_3

17   distributions, but the higher resolution highlights the fine-scale structures. During this flight a front passed

18   over LA region. Figure 10a shows that the low-altitude winds came from the northwest, and turned

19   southwesterly after passing over LA. In the upwind areas, west of longitude 119°W, the coarse and

20   nesting simulations did not show significant differences for O_3 and CO. Over and downwind of LA the

21   nesting simulated CO showed two high zones (Figure 10a), while the coarse simulation had only one

22   broad high-CO region (Figure 10b). The CO distributions mainly reflect the local emissions and transport

23   effects. The elevated O_3 regions in the nested-grid simulated did not overlap with the peak values of CO

24   (as was the case for the coarse simulation), but instead were shifted downwind. The region of high ozone

1   (> 120 ppbv) was significantly smaller on the fine grid, indicating the effect of model resolution on

2   photochemical processes.

3

4   To highlight the differences between these two simulations the flight segment shown in Figure 10c, where

5   the WP-3 aircraft operated around in and around LA, was analyzed further. Figure 11 shows these two

6   simulations compared to the aircraft measurements. The simulated ethyne and propane concentrations

7   using the nested grid were in much better agreement with the observations (Figure 11a, 11b). This is

8   mainly due to the better resolution of the emissions, which enables a better representation of peak values.

9   The refined high-resolution emissions better resolve the differences among urban, suburban and rural

10 areas.

11

12 The lifetime of CO is much longer than ethyne or propane, and it also has a higher background

13 concentrations and stronger emissions. Thus CO concentrations reflect the impacts of both local

14 emissions and transport. Both the coarse and fine grid simulated CO capture the main altitude variations

15 (Figure 11c), but the nested grid better represents the CO fine structures. Figure 11c also shows the

16 simulated CO chemical net budgets, which vary in a manner consistent with that for the CO

17 concentrations. The CO chemical budget is determined by the oxidations of CO and hydrocarbons. In the

18 low-altitude areas around LA hydrocarbons were elevated and CO had a positive net chemical budget that

19 was positively correlated with the hydrocarbon concentrations. Most primary hydrocarbons tend to be co-

20 emitted with CO in this region, and thus high CO concentrations usually correspond to high hydrocarbon

21 concentrations (Figures 11a, 11b, 11c). This is the reason why the CO net budget was correlated with CO

22 concentration. For this flight the CO positive chemical budget helped to maintain the high CO

23 concentration, but the high CO was mainly caused by strong emissions. The contribution of the CO

24 chemical budget to CO concentrations was relatively weak, less than 2.5 ppbv/hour (Figure 11c).

25

1   Both the fine and coarse grid simulations tended to overestimate O_3 concentration, but the nested one

2   shows better agreement with the observations (Figure 11d). The net O_3 chemical budget for most of the

3   flight segments was positive, up to 25 ppbv/hour. For the flight segment after 21:40 GMT the O_3

4   concentrations followed the variations in the O_3 chemical net budgets, implying that O_3 production was

5   the dominant influencing factor for O_3 concentrations (as expected). The O_3 chemical budgets on the fine

6   grid show larger absolute values and variation than for the coarse grid, consistent with the better ability to

7   resolve plume-like structures using the finer grid. The 12 km simulation improved the O_3 prediction, but

8   even finer grids are needed.

9

10   The improvement in ozone prediction with finer grids can be explained by looking at the predicted and

11   observed NO_z (NO_y-NO_x) concentrations (Figure 11e). As the photochemical product of NO_x, NO_z is

12   usually highly correlated with O_3 (Trainer et al., 1993). The fine grid simulation shows a much better

13   agreement with the observations, and the coarse grid tends to overestimate NO_z. Figure 11e also shows

14   the O_3/NO_z ratios. Both simulations agree well with the observations for the flight segment before 22:00

15   GMT, but tend to underestimate this ratio (especially for the coarse simulation) by overestimating NO_z

16   between 22:00 GMT and 23:00 GMT. The OPE difference between these two simulations and the

17   observations reflect their different photochemical behaviors, which is also highlighted by their OH

18   differences (Figure 11f). Both simulations systematically overestimated OH, but the nested simulation

19   showed improvement for some flight segments (e.g., 22:30GMT to 23:30GMT). However, the OH

20   overestimation is not caused by overprediction of the J-values. Figures 11g and 11h show $J[O_3 \rightarrow O_2 + O^1D]$

21   and $J[NO_2]$ simulations compared to the observations. $J[O_3 \rightarrow O_2 + O^1D]$ is a key factor for clear-

22   background OH, since $O^1D + H_2O \rightarrow 2OH$ (R#19 in Table 2) is the primary natural source of OH. The

23   simulation agrees well with the observations for $J[O_3 \rightarrow O_2 + O^1D]$, but overestimates $J[NO_2]$. However, the

24     $J[NO_2]$ overestimation is much smaller than the OH or O_3 overestimations for the flight segments before

25   23:00GMT, indicating that OH overestimation may be associated with other process. The observed peak

of OH appeared downwind of LA at low-altitude (23:00GMT in Figure 11f) where the observed J-values

were very low (Figures 11g and 11h). This implies that other photochemical processes had stronger

influence on OH than the R#19 only in the polluted area, which was also discussed before (Figure 6).

Tang (2002) showed that nested simulation could have different photochemical relationship from the

corresponding coarse simulation. In this WP-3 flight segment, these two simulations showed different

responses to the same J-values, due to their different pollutant loadings. Figure 11 shows that the nested

simulation yielded better agreement with the observation than the coarse one for most flight segments,

implying the impacts of changing resolution on emitted species, secondary-generated species and

photochemical system.

7. Mission-Wide Summary

The aircraft observations provide a valuable data set upon which to evaluate the models capabilities to

simulate trace species distributions in the Eastern Pacific. A mission-wide summary of the model

predicted and observed values are shown in Table 3. Presented are mean values for all 3-minute averaged

observations and simulated values for the 13 WP-3 flights parsed into and averaged over three altitude

layers. Also shown are the correlation coefficients (R). The RAMS model driven by NCEP reanalysis data

accurately represented the meteorological variables as shown for pressure, wind speed and relative

humidity. In terms of trace gas species the highest mean concentrations are generally found below 1 km,

indicative of the fact that low altitude flights tended to sample air dominated by emissions from North

America. PAN, HNO_3, and acetone showed highest concentrations in the 1-3 km layer, and O_3 (peak

value >3km) are the exceptions. The model performed best in the lower altitudes, and for primary emitted

species such as CO, ethane, propane, NO_2, NO, and toluene. All species with the exception of SO_2 and

xylene are predicted within a factor of 2, and many within a factor of 1.2. The correlation coefficients for

all the listed species are greater than 0.67. The predictions for the 1-3 km region for SO_2, O_3, NO_z are

usually better than the <1km layer values in terms of mean values reflecting the fact that this layer is not

1   so heavily impacted by the very local sources. However R-values are lower. The predictions >3km are

2   markedly poorer, with R-values typically lower than 0.5. J-values are a notable exception, and are

3   predicted with appreciable skill at all altitudes. J-value predictions mainly reflected the simulations for

4   aerosol optical properties and clouds. $J[NO_2]$ is very sensitive to fractional cloud, and the correlation

5   coefficients for it are relatively low compared to predictions for other J-values.

6

7   The values presented in Table 3 are a mixture of North American and Asian emission influenced air

8   masses. To further evaluate the model performance the data was stratified using model-calculated Asian

9   ratios. The statistics for data points with Asian ratios <20% and >80% are presented in Tables 4 and 5,

10   respectively. At altitudes below 1 km, the model performance for air masses with Asian ratio < 20% was

11   better than that with Asian ratio > 80%, as represented by the R values for emitted short-lived species,

12   such as SO_2, NO, NO_2, and their products HNO_3, NO_z, and NO_y. This decrease in performance with

13   distance from the emission source areas is due to the accumulation of errors in the transport, chemistry

14   and removal processes. This performance difference for emitted short-lived species also existed at the

15   altitude 1-3km, but not as significant as that below 1km. At altitudes above 3km, the model did not show

16   any systematic differences between the high and low Asian-impact data sets.

17

18   Further insights into the model performance are shown in Figure 12, where the observed and calculated

19   CO distributions are presented. The observed CO distribution shows a high degree of variability with

20   values of CO > 160 ppbv found at all altitudes. The highest values however were found below 2km

21   around LA. The model shows the same basic behavior and is able to accurately capture the mean behavior

22   (as shown in Table 3 where the model mean values are within 6% of the observations at all altitudes). The

23   distribution of the bias (simulated-observed) is presented in Figure 12c. The estimated VOC age is shown

24   in Figure 12d, and reveals that aged air is found at all attitudes and latitudes. No simple relationship

25   appears between CO bias and air mass age. Similar results are found for observed and modeled NO_y.

8. Trinidad Head Surface Site

During the ITCT-2K2 experiment extensive measurements were also performed at the Trinidad Head

ground station (41.05N, 124.15W) located on the coast of Northern California (Figure 1). The regional

model was also used to provide forecast support and post-analysis interpretation to the surface

observations. Figure 13 shows the model simulations compared to the surface measurements from Julian

day 110 to 143 (April 10 to May 13). Figure 13a shows that simulated wind speeds agreed with the

observed diurnal peak wind velocities, and reflected their daily variation trend. However, due to the

relatively coarse resolution the RAMS prediction failed to represent the nighttime calm winds at Trinidad

Head. Both the RAMS prediction and the observations showed that north and northeast winds prevailed

(Figure 13b). The simulated CO and NO_y concentrations are consistent with the observation as shown in

Figures 13c and 13d. Both the observed and the simulated values show that CO and NO_y are generally

correlated. These two plots also show the CFORS simulated anthropogenic CO Asian ratio and NO_x age.

In general CO Asian ratios >0.5 are associated with air masses older than 50 hours. The highest NO_y

values occurred for air masses older than ~50 hours. For example the CO and NO_y peak concentrations on

Julian day 113, 119-120 and 132 were associated with low surface wind speeds, implying that these peaks

were mainly due to the accumulation of pollutants emitted from surrounding sources and transported to

Trinidad Head under the low-wind speed situations.

Further insights into the role of local emission sources on Trinidad head is found in Figure 13e, where the

simulated and observed propane, CFORS simulated VOC age, and observed methyl-t-butyl ether

(MTBE), a compound mainly used in North America, are plotted. As shown the simulation results

accurately capture the observed propane features. MTBE is a good tracer for local sources, as shown by

its plume-like structure (Goldstein et al., this issue; Millet et al., this issue). The MTBE peaks correlate

with peaks in propane (and other pollutants), but propane shows broader maxima, reflecting the longer

1   lifetime of propane. The simulated VOC age shows clearly a reversed variation with observed MTBE,

2   providing further evidence for the utility of the VOC age indicator in helping to identify local and distant

3   sources. As shown by the CO results in Figure 13c, the fresh pollution events at Trinidad Head are

4   superimposed on air that is a mixture of Asian and American sources.

5

6   Both observed and simulated CO and propane show a clear declining trend during this one-month period

7   (Figures 13c and 13e). The CO background concentration declined from 150 ppbv in April to 125 ppbv in

8   May, while the propane background concentration declined from 0.3 to 0.15 ppbv. This seasonal variation

9   trend reflects the increased CO and VOC consumptions by OH as the sunshine time became longer and

10   photochemical reactions became more active (simulated daytime OH increased by about 30% during this

11   period).

12

13   There were two instruments measuring aerosol ions at Trinidad Head: PILS (particle-into-liquid sampler)

14   and the aerodyne AMS (Aerosol Mass Spectrometer, Allan et al., this issue). Both of these instruments

15   report a cut-off diameter of ~1μm to capture submicron ions. The measured and simulated aerosol sulfate,

16   nitrate and ammonium concentrations are presented in Figures 13f, 13g and 13h, respectively. As

17   discussed in Allan et al. (this issue), the AMS instrument observed consistently more nitrate than the PILS

18   instrument. This may be due to the presence of organic nitrates and/or amines. The simulated nitrate

19   values are shown in Figure 13g for <1μm and <2.5μm diameters. The model results show an appreciable

20   amount of nitrate in particles with diameters between 1μm to 2.5μm. The simulation curves for <1μm and

21   <2.5μm tend to encompass most of the observed nitrate, suggesting that differences in inlet cut-offs could

22   account for some of the differences between the measurements. In fact, the AMS 1μm cut-off diameter

23   was not absolute, and the measurements included some particles between 1 and 2μm. A second AMS

24   instrument configured to study size resolved aerosol composition did identify sub- as well as super-

25   micron nitrate, with the super-micron nitrate associated with sea salt particles. The sea salt mode is mainly

1   in the super-micron regime but the tail did go down into the submicron. The simulated fraction of nitrate

2   in the super-micron particles varied from 5% to ~100%, with a mean value of 30%, and with the super

3   micron fraction dominating under elevated sea salt conditions. Sea salt nitrate was also linked to a

4   suppression of chloride relative to sodium in the PILS observations as well as in the simulations.

5

6   The simulated and observed ammonium values are shown in Figure 13h. The AMS and PILS

7   measurements quantitatively agree (slope AMS versus PILS of 1.099, with r^2 of 0.58, Allan et al., this

8   issue)). The model predictions are generally consistent with the observations, with essentially all of the

9   ammonium in the particles below < 1μm, which was also verified by the AMS size-resolved

10   measurements. Our prediction overestimated fine sea salt. After we removing this overestimation,

11   ammonium prediction looks reasonable (Figure 13h).

12

13   **9. Conclusion**

14   In this paper a multi-scale modeling system comprised of a regional tracer model, regional chemical

15   transport model with a nested grid, and with lateral boundary conditions from the global model

16   MOZART, was used to analyze aircraft and surface observations obtained during the ITCT-2K2

17   experiment. This model system was shown to have substantive predictive capability for the region and

18   times of the NOAA WP-3 flights during the ITCT-2K2 experiment.

19

20   The model estimated air mass source and chemical age indicators were used to help analyze the

21   observations. The Asian air masses were found to have lower ΔAcetone/ΔCO, ΔMethanol/ΔCO, and

22   ΔPropane/ΔEthyne ratios than air masses impacted by American sources. During the long-distance

23   transport Asian air masses experienced significant NO_y losses, associated with gas-aerosol interaction and

24   removal processes. The VOC and NO_x age analysis indicated that air masses measured had chemical ages

25   ranging from <0.5 to > 10 days. Asian air masses were usually aged (> 5 days). CO and O_3 were found to

have net negative budgets in the places dominated by Asian influences, with CO losses due to reaction with OH, and O_3 lost via photolytic dissociation. In the MBL OH levels were elevated and CO destruction was high. CO and O_3 chemical net budgets turned positive in polluted American air masses.

The aerosol partitioning of nitrate was shown to be important. Without aerosol consideration, the model overestimated HNO_3, especially for its background concentration along the U.S west coast. This result also implies that NO_y should decrease because of this process during the trans-Pacific transport.

Several big cities exist along the US west coast. This study indicated that these cities received little Asian influence, and that photochemical processes were dominated by local sources. At Trinidad Head, most high concentrations of pollutants are associated with calm-wind scenarios, and the accumulation of local sources. Seasonal variations of CO and VOCs showed a clear declining trend from April to May at this site. The high-resolution nested simulation greatly improved the prediction for emitted and secondary-generated species. The resolution change affected not only the emission intensity and distribution, but also the associated photochemical behavior. In the coarse resolution (60km) mode the concentrations of primary emitted species were underestimated, O_3 production was widespread, and the titrating effect of high-concentrated pollutants was not resolved. This resulted in an overestimation of O_3, NO_z and OH. The nesting simulation significantly improved the model performance.

The multi-scale modeling system consisting of nested models was shown to provide a viable tool for studying trans-Pacific transport, and the results were often consistent with observations. However, the difference between the simulations and measurements indicate that further improvements are needed to better quantify multi-scale and multi-source problems.

Acknowledgement:

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Table 1. The 17 vertical layers (over sea surface) used in CFORS and STEM-2K3 and CO background profile in the Eastern Pacific

Altitude(km)	0.075	0.24	0.44	0.68	0.96	1.3	1.7	2.2	2.8	3.5	4.4	5.4	6.6	8.1	9.8	11.6	13.4
CO (ppbv)	120	120	120	120	120	120	120	110	110	110	110	108	105	105	98	85	80

Table 2. Reaction index as mentioned in Figure 6

Reaction Index #	Reaction Equations
2	$O^3P + O_2 \rightarrow O_3$
7	$O_3 + NO \rightarrow NO_2 + O_2$
17	$O_3 + hv \rightarrow O_2 + O^3P$
18	$O_3 + hv \rightarrow O_2 + O^1D$
19	$O^1D + H_2O \rightarrow 2OH$
21	$OH + NO \rightarrow HONO$
22	$HONO + hv \rightarrow OH + NO$
25	$OH + NO_2 \rightarrow HNO_3$
29	$OH + CO + O_2 \rightarrow HO_2 + CO_2$
30	$OH + O_3 \rightarrow HO_2 + O_2$
31	$HO_2 + NO \rightarrow NO_2 + OH$
32	$HO_2 + NO_2 \rightarrow HNO_4$
33	$HNO_4 + hv \rightarrow 0.61HO_2 + 0.61NO_2 + 0.39OH + 0.39NO_3$
36	$HO_2 + O_3 \rightarrow OH + 2O_2$
37	$HO_2 + HO_2 \rightarrow H_2O_2 + O_2$
38	$HO_2 + HO_2 + H_2O \rightarrow H_2O_2 + O_2 + H_2O$
41	$H_2O_2 + hv \rightarrow 2OH$
42	$H_2O_2 + OH \rightarrow HO_2 + H_2O$
43	$HO_2 + OH \rightarrow O_2 + H_2O$
44	$OH + SO_2 + H_2O + O_2 \rightarrow H_2SO_4 + HO_2$
45	$H_2 + OH \rightarrow HO_2$
46	$CH_3O_2 + NO + O_2 \rightarrow NO_2 + HCHO + HO_2$
51	$RO_2\_R + NO \rightarrow NO_2 + HO_2$
123	$HCHO + hv + 2O_2 \rightarrow 2HO_2 + CO$
124	$HCHO + hv \rightarrow CO + H_2$
125	$OH + HCHO \rightarrow HO_2 + CO$
126	$HCHO + HO_2 \rightarrow HOCOO$
127	$HOCOO \rightarrow HCHO + HO_2$
130	$CH_3CHO + OH \rightarrow CCO\_O_2$
131	$CH_3CHO + hv \rightarrow CO + HO_2 + CH_3O_2$
133	$RCHO + OH \rightarrow 0.034CO + 0.034CH_3CHO + Other$
134	$RCHO + hv \rightarrow CH_3CHO + CO + HO_2 + RO_2\_R$
200	$C_2H_2 + OH \rightarrow 0.603OH + 0.297HO_2 + 0.393CO + 0.1RO_2\_R + 0.096HCHO$

*RCHO is lumped aldehydes with three and more carbons, RO_2_R is peroxy radical operator representing NO to NO_2 conversion with HO_2 formation, and CCO_O_2 represents acetyl peroxy radicals (Carter, 2000)

Table 3. Observed and Simulated (60 km resolution) Mean Values and Their Correlation Coefficients (R) for ITCT-2K2 all WP-3 flights

Species and Variables	Below 1km			1km to 3km			Above 3km		
	Observed	Modeled	R	Observed	Modeled	R	Observed	Modeled	R
Pressure (hPa)	971.5	963.1	0.96	801.2	798.5	0.99	554.4	551.7	0.99
Wind Speed (m/s)	8.2	8.4	0.67	8.0	7.3	0.68	12.7	11.5	0.88
Relative Humidity (%)	68.8	58.3	0.82	28.8	33.9	0.85	27.5	28.2	0.73
CO (ppbv)	154	163	0.92	135	147	0.60	120	113	0.18
O_3 (ppbv)	47.2	58.1	0.81	57.3	57.8	0.66	59.8	64.0	0.47
Ethane (ppbv)	1.62	1.52	0.82	1.31	1.15	0.22	0.98	0.77	0.40
Propane (ppbv)	0.60	0.36	0.83	0.32	0.22	0.34	0.13	0.06	0.30
Ethyne (ppbv)	0.33	0.34	0.86	0.23	0.23	0.47	0.16	0.11	0.14
SO_2 (ppbv)	0.76	0.23	0.52	0.44	0.35	0.41	0.30	0.10	0.18
Acetone (ppbv)	1.03	0.71	0.89	1.20	0.55	0.64	0.89	0.27	0.28
PAN (ppbv)	0.16	0.57	0.68	0.44	0.47	0.06	0.14	0.15	0.26
NO_2 (ppbv)	1.76	1.04	0.88	0.58	0.49	0.65	0.055	0.018	0.43
NO (ppbv)	0.52	0.31	0.83	0.15	0.13	0.69	0.019	0.009	0.43
HNO_3 (ppbv)	0.80	0.65	0.87	0.81	0.64	0.73	0.13	0.12	0.60
NO_y (ppbv)	3.3	3.8	0.92	1.75	2.54	0.66	0.42	0.42	0.51
NO_z (ppbv)	1.64	2.46	0.95	1.25	1.92	0.66	0.38	0.40	0.57
Toluene (ppbv)*	0.113	0.102	0.91	0.0492	0.0604	0.69	0.019	0.017	0.18
Xylene (ppbv)*	0.150	0.051	0.85	0.0279	0.0195	0.73	0.0047	0.0035	0.20
$J[NO_2]$ (1/s)	6.66×10^{-3}	7.62×10^{-3}	0.67	9.1×10^{-3}	8.83×10^{-3}	0.58	0.010	0.0092	0.56
$J[O_3{\rightarrow}O_2+O^1D]$ (1/s)	2.00×10^{-5}	2.30×10^{-5}	0.86	2.92×10^{-5}	2.77×10^{-5}	0.84	2.95×10^{-5}	2.77×10^{-5}	0.83
$J[H_2O_2]$ (1/s)	4.31×10^{-6}	5.75×10^{-6}	0.79	6.21×10^{-6}	6.88×10^{-6}	0.74	6.72×10^{-6}	7.10×10^{-6}	0.77
$J[HNO_3]$ (1/s)	3.67×10^{-7}	4.94×10^{-7}	0.84	5.34×10^{-7}	5.93×10^{-7}	0.81	5.47×10^{-7}	5.90×10^{-7}	0.83
$J[HNO_2{\rightarrow}OH+NO]$ (1/s)	1.46×10^{-3}	1.49×10^{-3}	0.68	2.02×10^{-3}	1.73×10^{-3}	0.60	2.26×10^{-3}	1.81×10^{-3}	0.59
$J[HCHO{\rightarrow}H+HCO]$ (1/s)	1.87×10^{-5}	2.49×10^{-5}	0.81	2.75×10^{-5}	3.05×10^{-5}	0.77	3.20×10^{-5}	3.37×10^{-5}	0.79
$J[HCHO{\rightarrow}H_2+CO]$ (1/s)	3.13×10^{-5}	3.58×10^{-5}	0.74	4.65×10^{-5}	4.33×10^{-5}	0.69	5.58×10^{-5}	4.70×10^{-5}	0.69

* measured with proton-transfer-reaction mass spectrometry (PTR-MS).

Table 4. Observed and Simulated (60 km resolution) Mean Values and Their Correlation Coefficients (R) for ITCT-2K2 WP-3 flights with CO Asian Ratio < 20%

Species and Variables	Below 1km			1km to 3km			Above 3km		
	Observed	Modeled	R	Observed	Modeled	R	Observed	Modeled	R
CO (ppbv)	172	196	0.91	146	177	0.60	110	119	0.37
O_3 (ppbv)	51.8	71.6	0.81	61.8	67.8	0.71	50. 2	54.3	0.54
Ethane (ppbv)	1.86	1.90	0.82	1.43	1.41	0.16	0.93	0.73	0.27
Propane (ppbv)	0.89	0.54	0.83	0.51	0.35	0.27	0.21	0.09	0.26
Ethyne (ppbv)	0.45	0.41	0.87	0.28	0.27	0.46	0.18	0.09	0.22
SO_2 (ppbv)	1.16	0.40	0.48	0.70	0.64	0.31	0.26	0.16	-0.02
Acetone (ppbv)	1.42	0.91	0.89	1.53	0.73	0.56	0.99	0.34	0.35
PAN (ppbv)	0.23	0.96	0.60	0.79	0.85	-0.10	0.12	0.25	0.40
NO_2 (ppbv)	2.65	1.90	0.88	1.15	1.01	0.61	0.17	0.10	0.34
NO (ppbv)	0.89	0.57	0.82	0.31	0.27	0.65	0.06	0.04	0.35
HNO_3 (ppbv)	1.39	1.17	0.86	1.58	1.28	0.68	0.33	0.36	0.73
NO_y (ppbv)	5.45	6.78	0.92	3.38	4.98	0.61	0.78	1.11	0.47
NO_z (ppbv)	2.40	4.31	0.95	2.13	3.69	0.60	0.61	0.98	0.56
Toluene (ppbv)	0.156	0.174	0.91	0.069	0.109	0.67	0.029	0.017	0.25
Xylene (ppbv)	0.165	0.091	0.84	0.034	0.039	0.72	0.004	0.003	0.26
$J[NO_2]$ (1/s)	6.53×10^{-3}	6.90×10^{-3}	0.93	8.11×10^{-3}	7.86×10^{-3}	0.44	0.0084	0.0050	0.65
$J[O_3\rightarrow O_2+O^1D]$ (1/s)	2.11×10^{-5}	2.07×10^{-5}	0.98	2.78×10^{-5}	2.29×10^{-5}	0.92	3.22×10^{-5}	1.38×10^{-5}	0.96
$J[H_2O_2]$ (1/s)	4.31×10^{-6}	5.21×10^{-6}	0.97	5.61×10^{-6}	5.99×10^{-6}	0.68	6.08×10^{-6}	3.71×10^{-6}	0.89
$J[HNO_3]$ (1/s)	3.77×10^{-7}	4.47×10^{-7}	0.98	4.96×10^{-7}	5.08×10^{-7}	0.84	5.53×10^{-7}	3.05×10^{-7}	0.94
$J[HNO_2\rightarrow OH+NO]$ (1/s)	1.43×10^{-3}	1.34×10^{-3}	0.94	1.79×10^{-3}	1.53×10^{-3}	0.46	1.88×10^{-3}	0.98×10^{-3}	0.68
$J[HCHO\rightarrow H+HCO]$ (1/s)	1.86×10^{-5}	2.22×10^{-5}	0.98	2.45×10^{-5}	2.58×10^{-5}	0.72	2.76×10^{-5}	1.68×10^{-5}	0.90
$J[HCHO\rightarrow H_2+CO]$ (1/s)	3.11×10^{-5}	3.23×10^{-5}	0.96	4.13×10^{-5}	3.74×10^{-5}	0.58	4.57×10^{-5}	2.43×10^{-5}	0.79

Table 5. Observed and Simulated (60 km resolution) Mean Values and Their Correlation Coefficients (R) for ITCT-2K2 WP-3 flights with CO Asian Ratio > 80%

Species and Variables	Below 1km			1km to 3km			Above 3km		
	Observed	Modeled	R	Observed	Modeled	R	Observed	Modeled	R
CO (ppbv)	129	121	0.90	127	118	0.46	125	112	0.085
O_3 (ppbv)	40.7	38.1	0.74	53.2	47.5	0.31	64.4	66.9	0.46
Ethane (ppbv)	1.29	0.97	0.95	1.23	0.86	0.73	1.03	0.77	0.44
Propane (ppbv)	0.20	0.11	0.97	0.18	0.08	0.87	0.13	0.06	0.33
Ethyne (ppbv)	0.20	0.23	0.85	0.19	0.17	0.64	0.17	0.11	0.17
SO_2 (ppbv)	0.21	0.021	-0.02	0.21	0.09	0.35	0.32	0.11	0.26
Acetone (ppbv)	0.48	0.43	0.77	0.86	0.35	-0.18	0.92	0.25	0.18
PAN (ppbv)	0.060	0.075	0.77	0.089	0.094	0.76	0.15	0.14	0.33
NO_2 (ppbv)	0.25	0.020	0.12	0.035	0.011	-0.01	0.026	0.006	0.30
NO (ppbv)	0.074	0.008	0.28	0.011	0.006	0.19	0.008	0.005	0.18
HNO_3 (ppbv)	0.077	0.058	-0.35	0.070	0.053	0.28	0.058	0.069	0.25
NO_y (ppbv)	0.40	0.24	0.30	0.30	0.23	0.38	0.34	0.29	0.31
NO_z (ppbv)	0.23	0.21	0.007	0.26	0.22	0.53	0.34	0.28	0.31
$J[NO_2]$ (1/s)	7.53×10^{-3}	9.23×10^{-3}	0.31	9.88×10^{-3}	1.03×10^{-2}	0.61	0.0101	0.0104	0.63
$J[O_3 \rightarrow O_2 + O^1D]$ (1/s)	1.89×10^{-5}	3.02×10^{-5}	0.72	3.23×10^{-5}	3.59×10^{-5}	0.81	2.88×10^{-5}	3.34×10^{-5}	0.83
$J[H_2O_2]$ (1/s)	4.63×10^{-6}	7.13×10^{-6}	0.54	6.82×10^{-6}	8.30×10^{-6}	0.75	6.70×10^{-6}	8.22×10^{-6}	0.83
$J[HNO_3]$ (1/s)	3.70×10^{-7}	6.26×10^{-7}	0.66	5.87×10^{-7}	7.33×10^{-7}	0.80	5.40×10^{-7}	6.94×10^{-7}	0.85
$J[HNO_2 \rightarrow OH + NO]$ (1/s)	1.65×10^{-3}	1.80×10^{-3}	0.34	2.20×10^{-3}	2.02×10^{-3}	0.62	2.25×10^{-3}	2.06×10^{-3}	0.66
$J[HCHO \rightarrow H + HCO]$ (1/s)	2.00×10^{-5}	3.16×10^{-5}	0.59	3.07×10^{-5}	3.78×10^{-5}	0.76	3.20×10^{-5}	3.94×10^{-5}	0.84
$J[HCHO \rightarrow H_2 + CO]$ (1/s)	3.45×10^{-5}	4.45×10^{-5}	0.47	5.12×10^{-5}	5.21×10^{-5}	0.70	5.60×10^{-5}	5.42×10^{-5}	0.76

Figure Captions

Figure 1. Model domains, NOAA WP-3 flight paths (colored lines), and estimated CO emissions on the various domains.

Figure 2.   ITCT aircraft-observed correlations (left panels) classified by simulated Asian Ratio < 20% and > 80%, and TRACE-P aircraft measured correlations (right panels). All plots are marked with linear fit lines and correlation coefficient R.

Figure 3. Observed ratios versus VOC age and NO_x age estimated by the CFORS model for all ITCT flights, color-coded by VOC age.

Figure 4. Simulated results at 21:00GMT, April 25, 2002. CO Asian ratios are presented in contour line in plots a and b. CO and O_3 budgets are in ppbv/hr.

Figure 5. STEM simulations compared to measurements for the WP-3 flight on April 25.

Figure 6. Simulated species budgets and their components for O_3, CO, OH, HO_2 and HCHO at 19:00 GMT (altitude 2837m), 19:03 GMT (altitude 3630m), and 21:00 GMT (altitude 470m) along the WP-3 flight path on April 25. Some net budgets, concentrations, and rate terms are multiplied a factor of 10 for display purposes. Negative values represent loss terms. For each term, the bar sequence from left to right is 19:00 GMT, 19:03 GMT, and 21:00 GMT. The reaction index (R#) is listed in Table 2.

Figure 7. Simulated results at 21:00GMT, May 15, 2002. CO Asian ratios are presented by contour line in panels a and b.

Figure 8. STEM simulations compared to measurements for the WP-3 flight on May 15.

Figure 9. STEM simulations with and without the aerosol consideration compared to observations of 4 WP-3 flights, whose paths are shown by the plot e.

Figure 10. Coarse (60km resolution) and nested (12 km resolution) simulated CO and O_3 at 400m at 21GMT, May 13, for the WP-3 flight 10 over Los Angeles and surrounding area.

Figure 11. Simulations compared to the measurements along the path of WP-3 flight 10 shown in Figure 12c. The nested and coarse simulations have similar J-value predictions. So plots g and h show only one simulation.

Figure 12. Latitude-Altitude distributions of CO, CO difference, VOC age, NO_y, and NO_y difference for All WP-3 flights.

Figure 13. Simulation compared to surface observations at Trinidad Head.
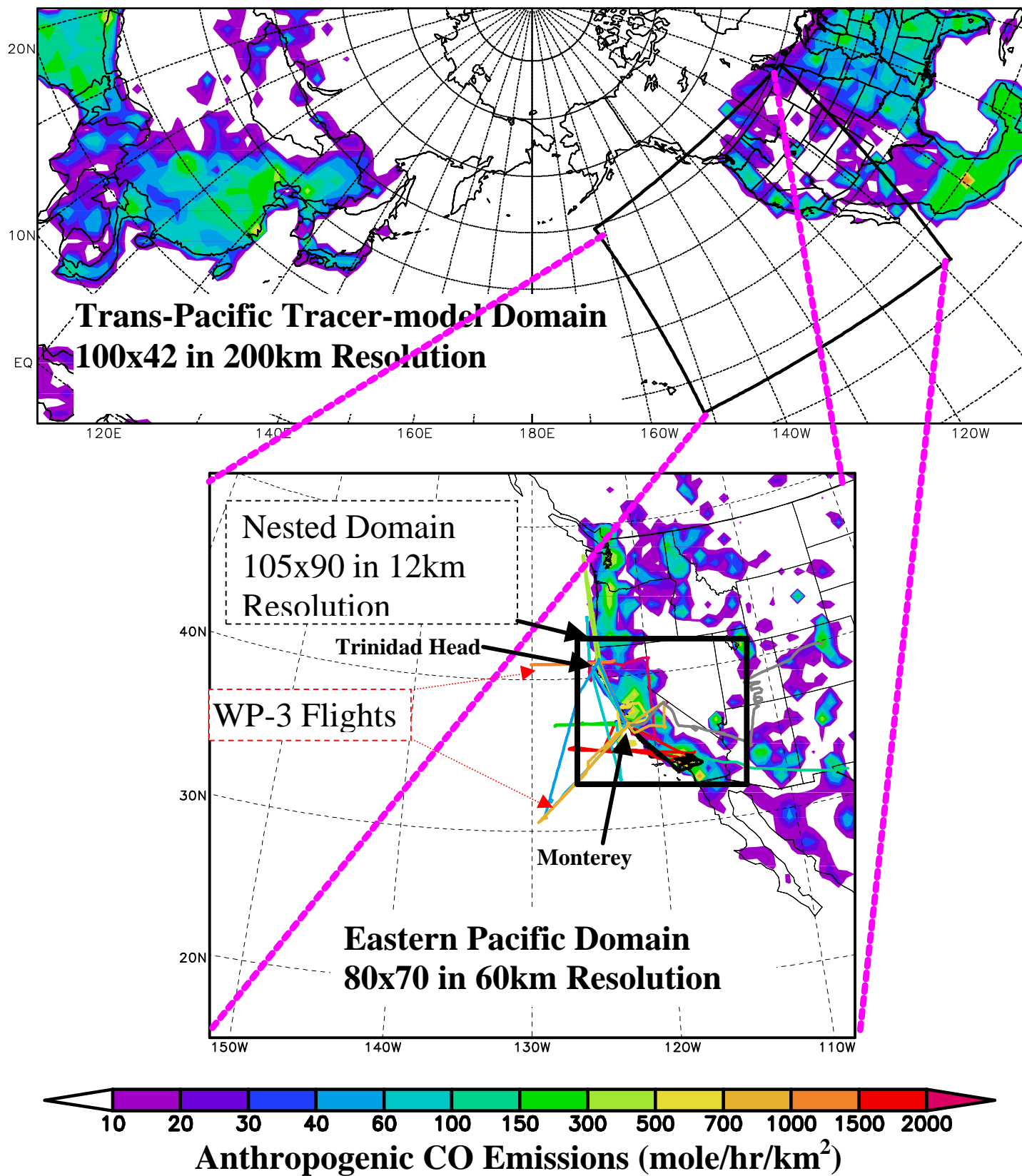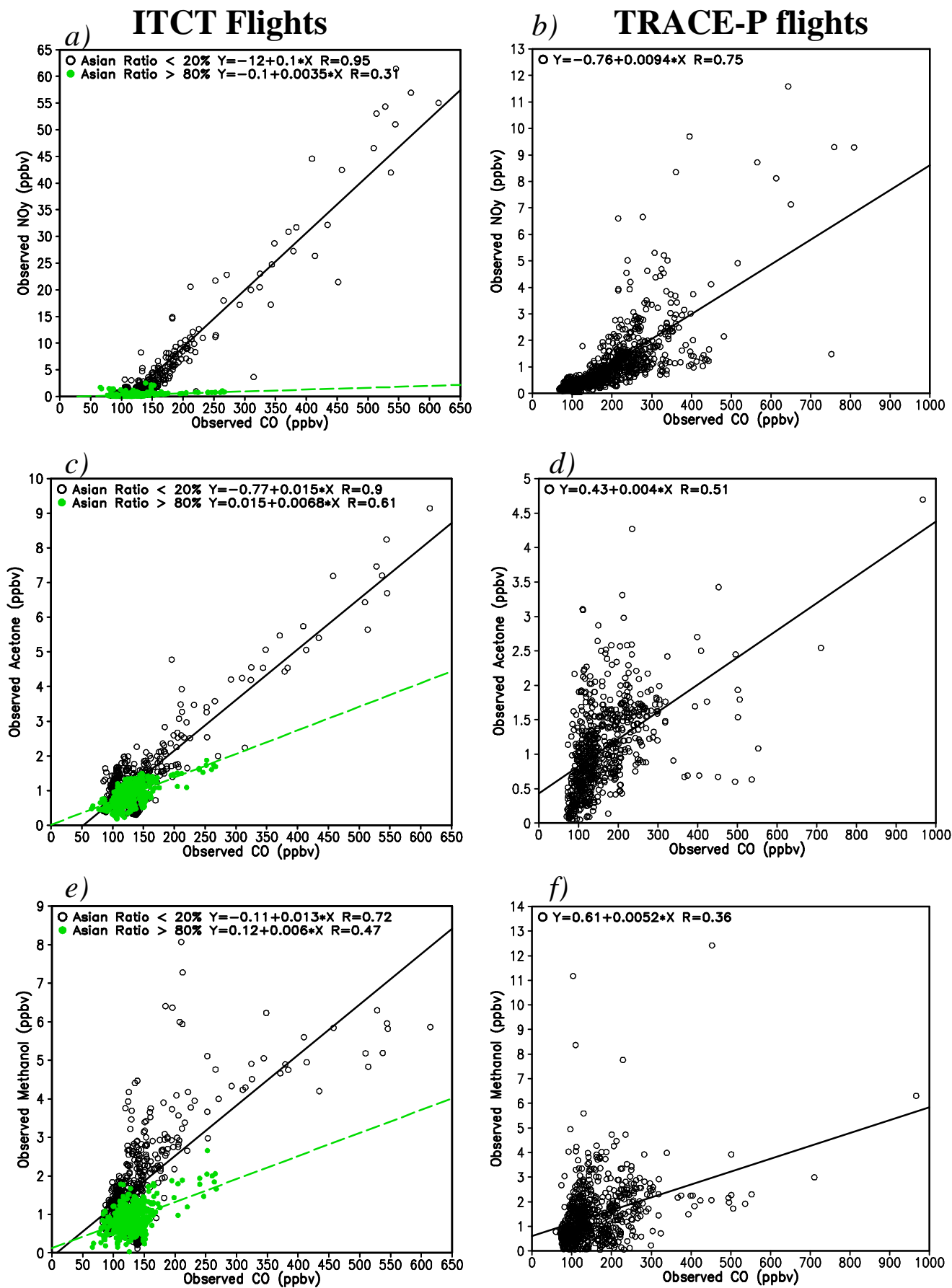
Trans-Pacific Tracer-model Domain 100x42 in 200km Resolution

Nested Domain 105x90 in 12km Resolution

Trinidad Head

WP-3 Flights

Monterey

Eastern Pacific Domain 80x70 in 60km Resolution

Anthropogenic CO Emissions (mole/hr/km^2)

10 20 30 40 60 100 150 300 500 700 1000 1500 2000

Figure 1

ITCT Flights

a)

○ Asian Ratio < 20% Y=−12+0.1*X R=0.95
● Asian Ratio > 80% Y=−0.1+0.0035*X R=0.31

c)

○ Asian Ratio < 20% Y=−0.77+0.015*X R=0.9
● Asian Ratio > 80% Y=0.015+0.0068*X R=0.61

e)

○ Asian Ratio < 20% Y=−0.11+0.013*X R=0.72
● Asian Ratio > 80% Y=0.12+0.006*X R=0.47

TRACE-P flights

b)

○ Y=−0.76+0.0094*X R=0.75

d)

○ Y=0.43+0.004*X R=0.51

f)

○ Y=0.61+0.0052*X R=0.36

Figure 2 (continued in next page)

Figure 2 (continued)

g)

Observed Propane (ppbv) vs Observed Ethyne (ppbv)

○ Asian Ratio < 20% Y=−0.096+2.1*X R=0.87
● Asian Ratio > 80% Y=−0.054+1.2*X R=0.88

h)

Observed Propane (ppbv) vs Observed Ethyne (ppbv)

○ Y=0.072+0.61*X R=0.78

i)

Observed O₃ (ppbv) vs Observed NOz (ppbv)

○ Asian Ratio < 20% Y=49+3.2*X R=0.70
● Asian Ratio > 80% Y=30+105*X R=0.80

i)

Observed O₃ (ppbv) vs Observed NOz (ppbv)

○ Y=54.4+4.1*X R=0.37

Figure 2

Figure 3

a) Simulated Anthropogenic CO and Its Asian Ratio in the 1 km layer

Tracer CO Concentrations (ppbv)

2 4 6 10 15 20 30 40 60 70 100 130

b) Simulated Anthropogenic CO and Its Asian Ratio in the 5.4 km layer

Tracer CO Concentrations (ppbv)

0.05 0.1 0.2 0.3 0.5 1 1.5 2 3 4 6 10 15

Figure 4

Figure 4 (continued)

c) **Simulated O₃ in the 1 km layer**
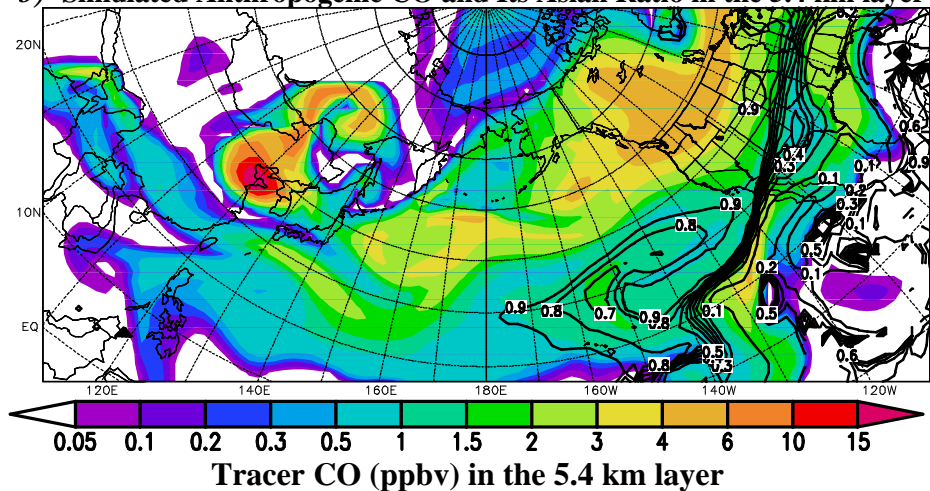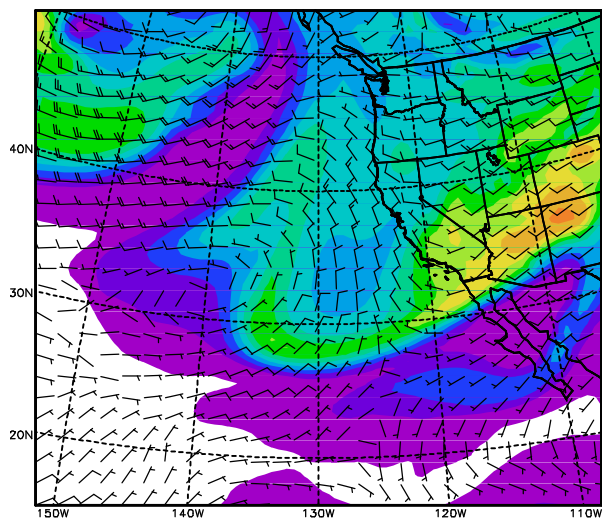
d) **Simulated O₃ in the 5.4 km layer**

e)

Simulated O₃ (color-coded) and its chemical net budget (contour) over latitude 35°N

Simulated O₃ (ppbv)

f)

Simulated CO (color-coded) and its chemical net budget (contour) over latitude 35°N

Simulated CO (ppbv)
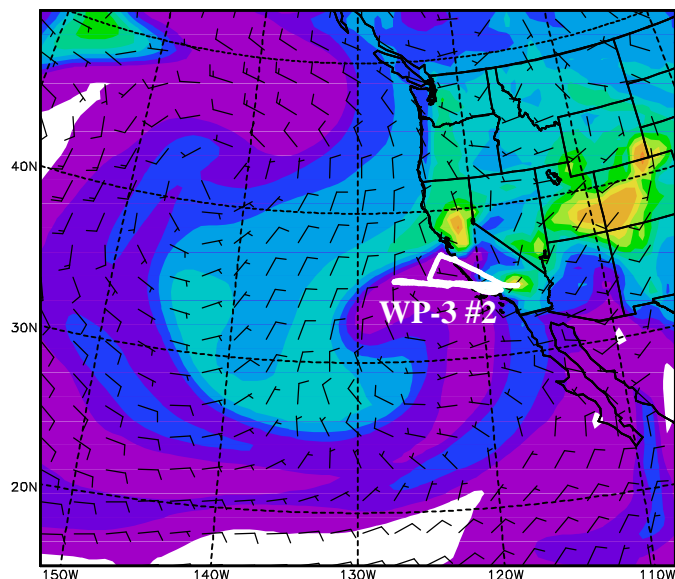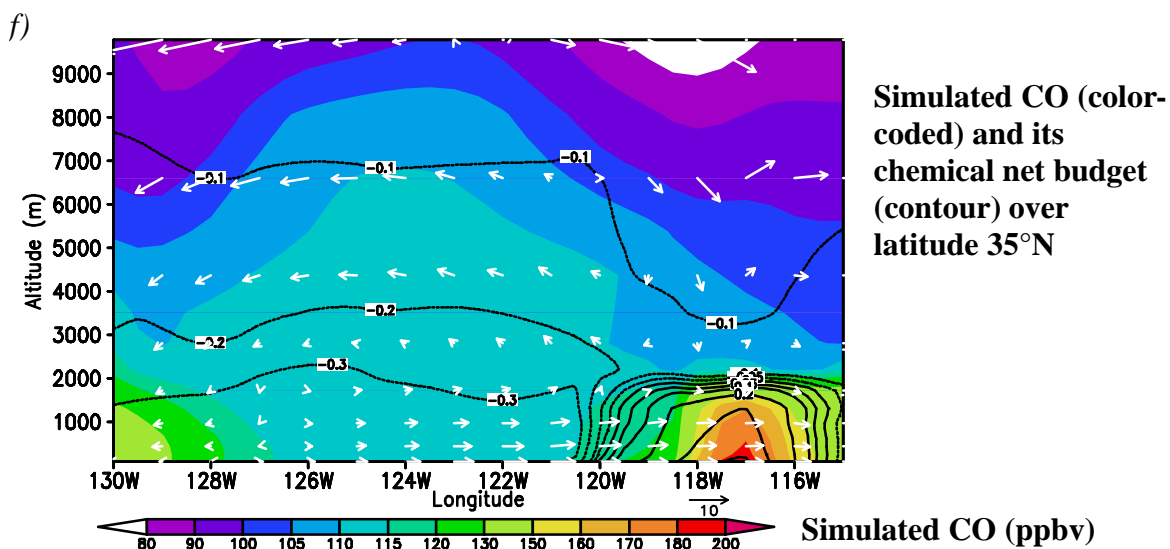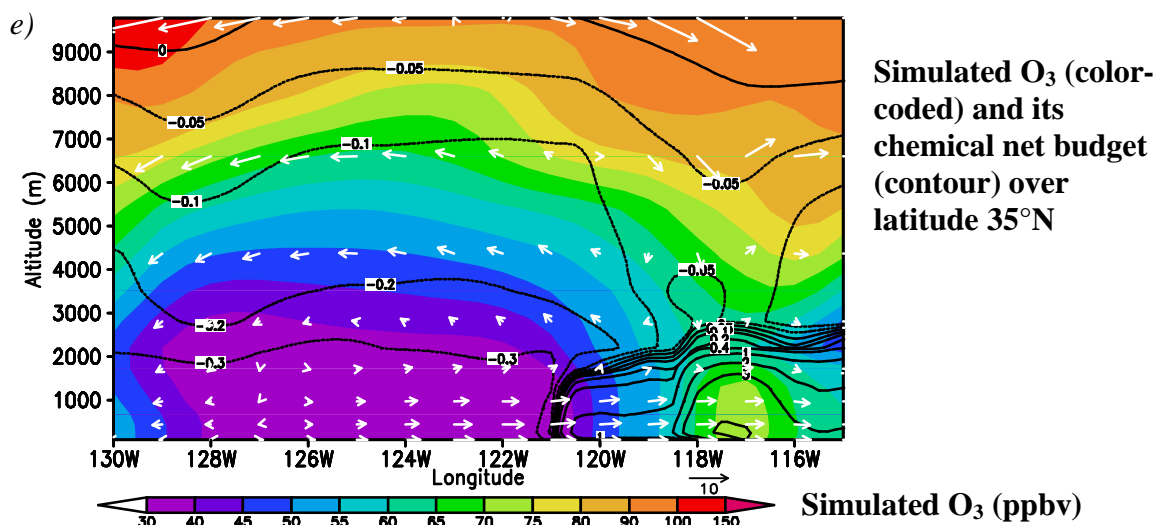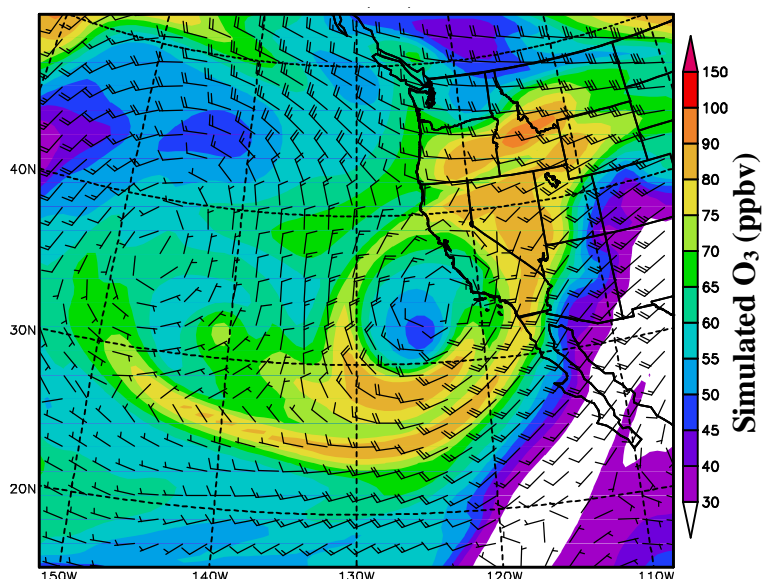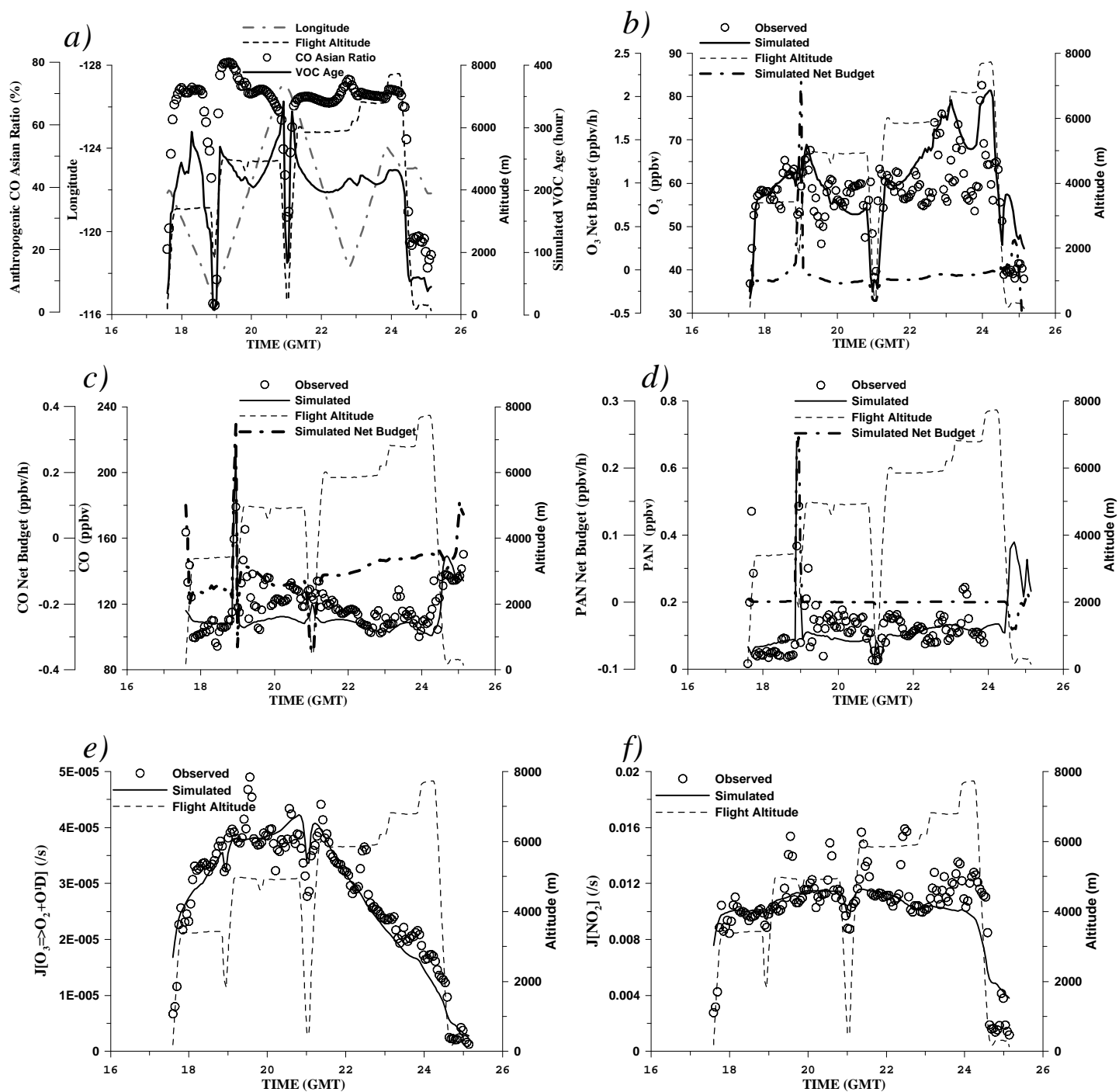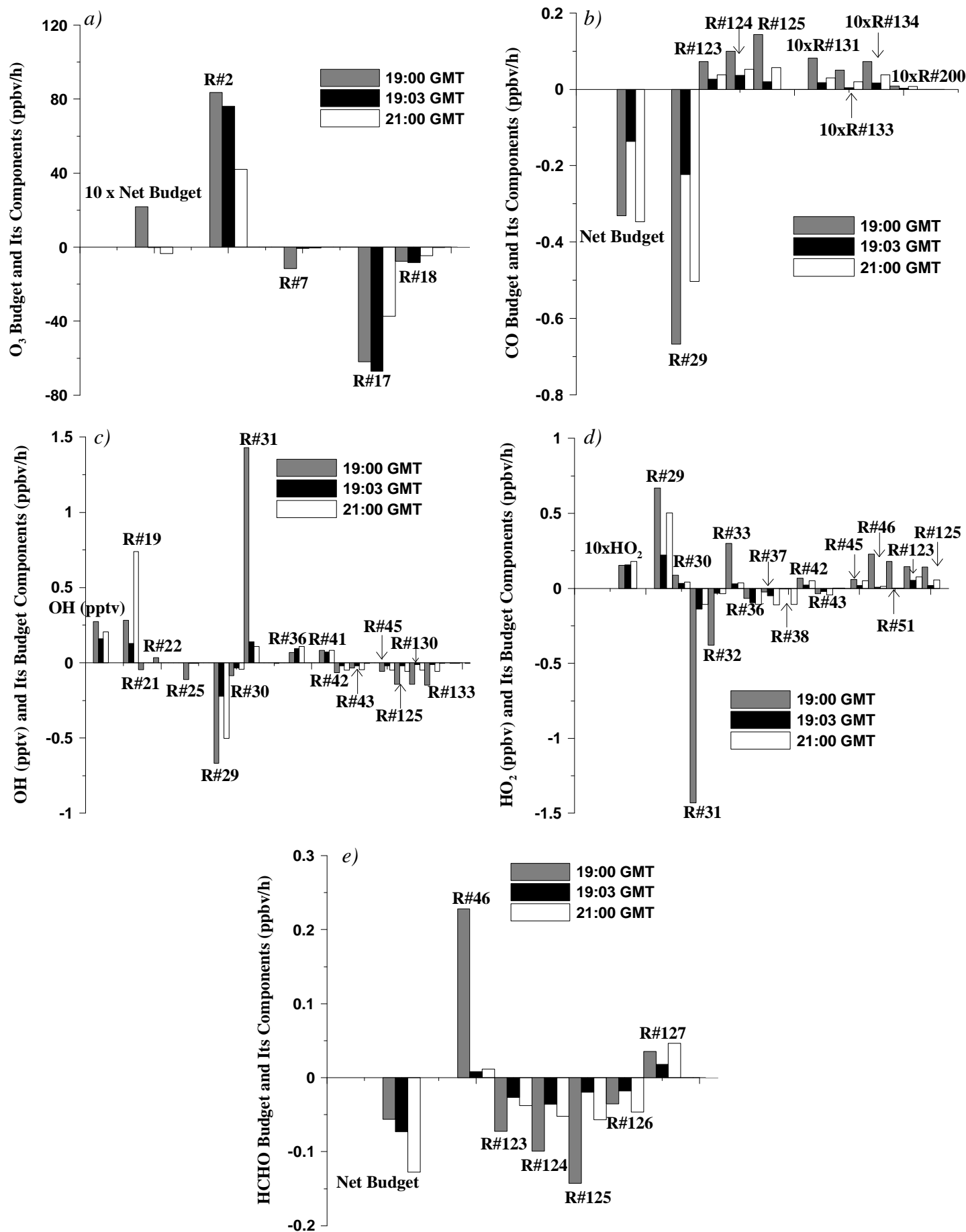
Figure 5

Figure 6

a) Simulated Anthropogenic CO and Its Asian Ratio in the 1 km layer

b) Simulated Anthropogenic CO and Its Asian Ratio in the 5.4 km layer

Tracer CO (ppbv) in the 1 km layer
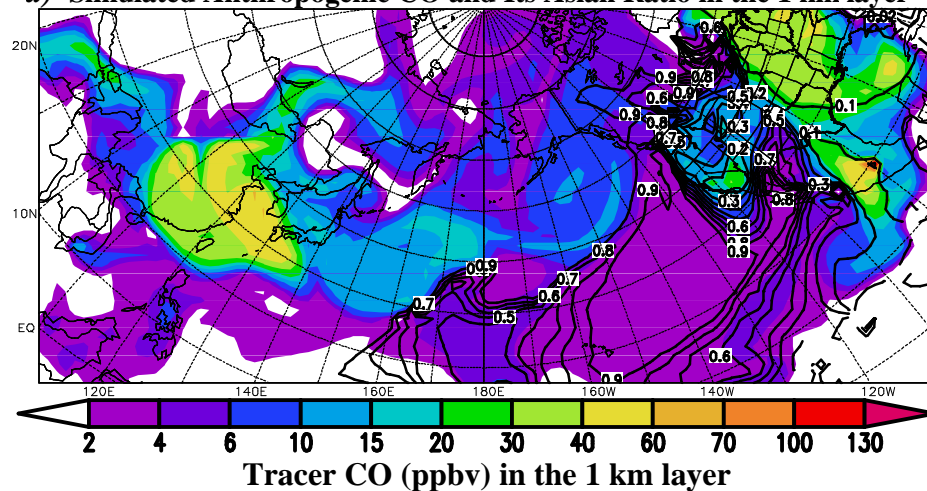
Tracer CO (ppbv) in the 5.4 km layer

c) Simulated O$_3$ in the 1 km layer
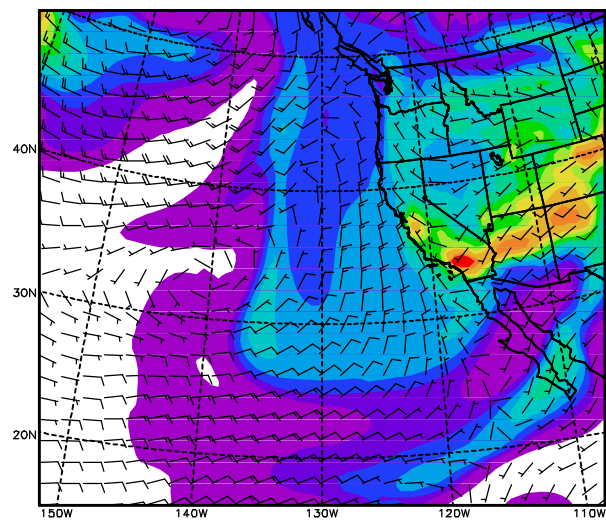
d) Simulated O$_3$ in the 2.8 km layer

e) Simulated O$_3$ in the 5.4 km layer

Simulated O$_3$ (ppbv)
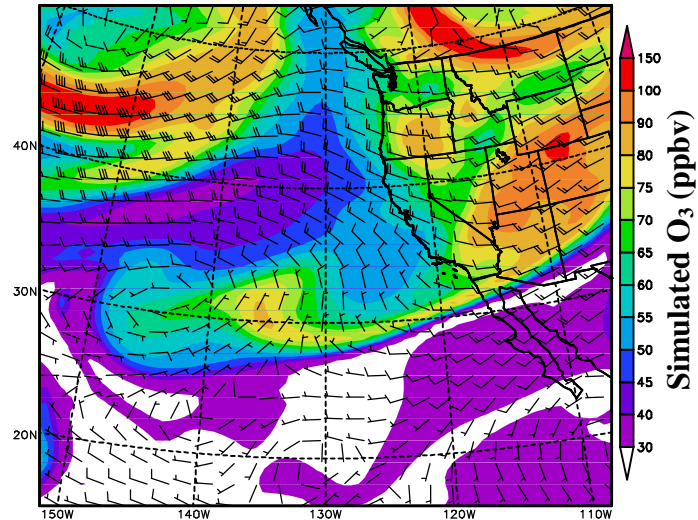
Figure 7

Figure 7 (continued)

f) **Simulated O$_3$ Budget in the 1 km layer**   *g)* **Simulated O$_3$ Budget in the 2.8 km layer**   *h)* **Simulated O$_3$ Budget in the 5.4 km layer**

WP-3 #11

i) **Simulated CO Budget in the 1 km layer**   *j)* **Simulated CO Budget in the 2.8 km layer**   *k)* **Simulated CO Budget in the 5.4 km layer**

Simulated O$_3$ Budget (ppbv/hr)

Simulated CO Budget (ppbv/hr)
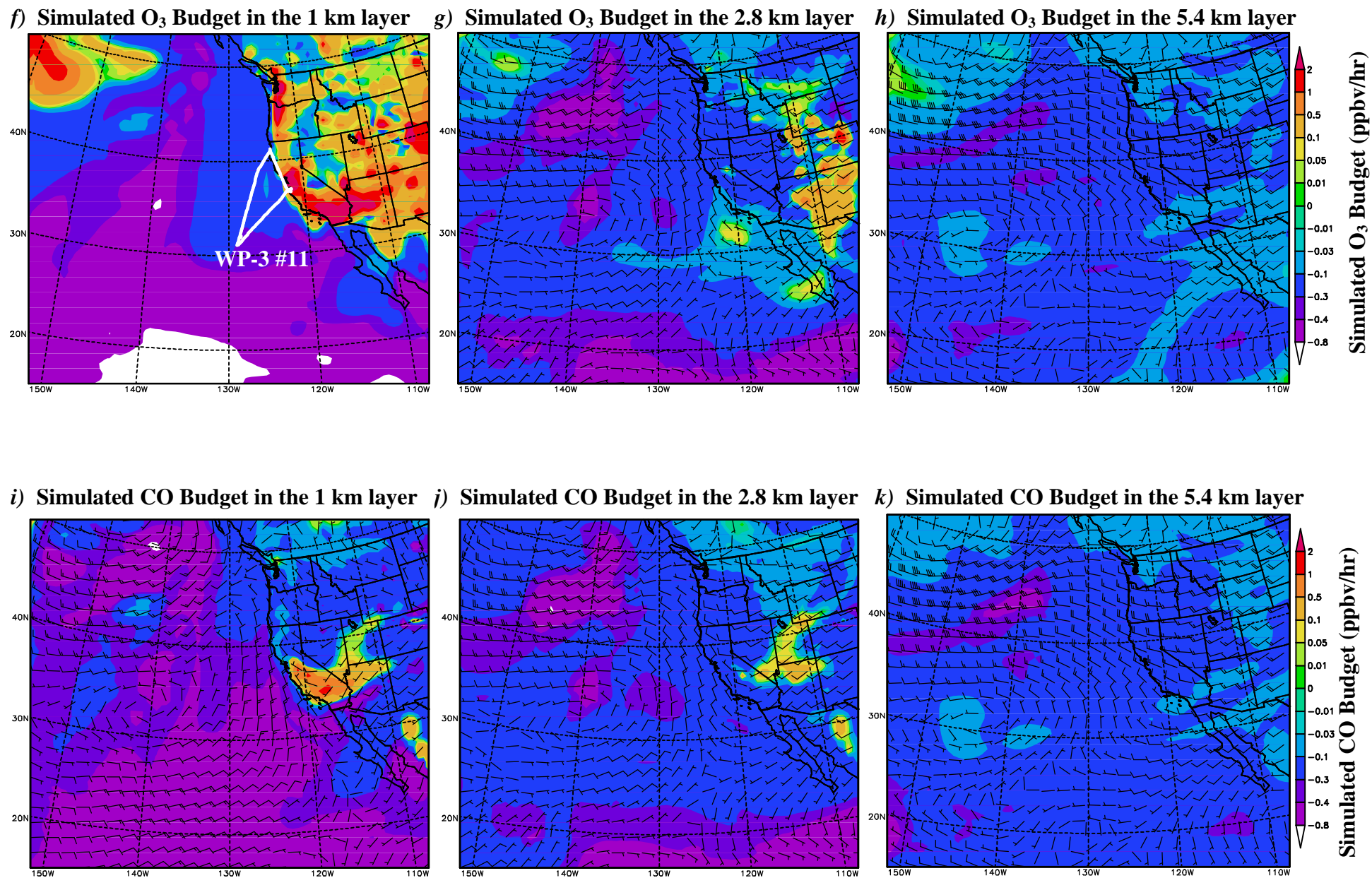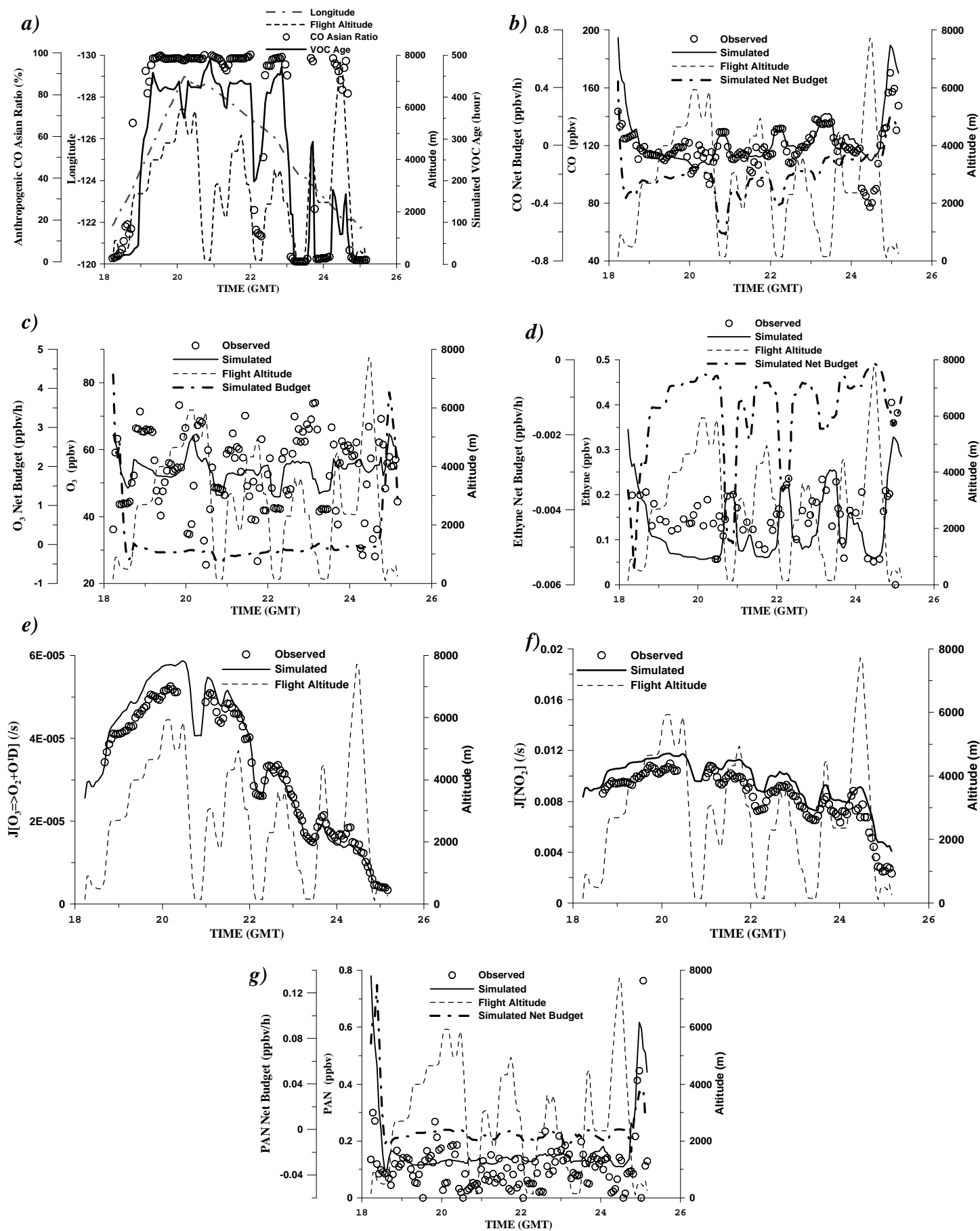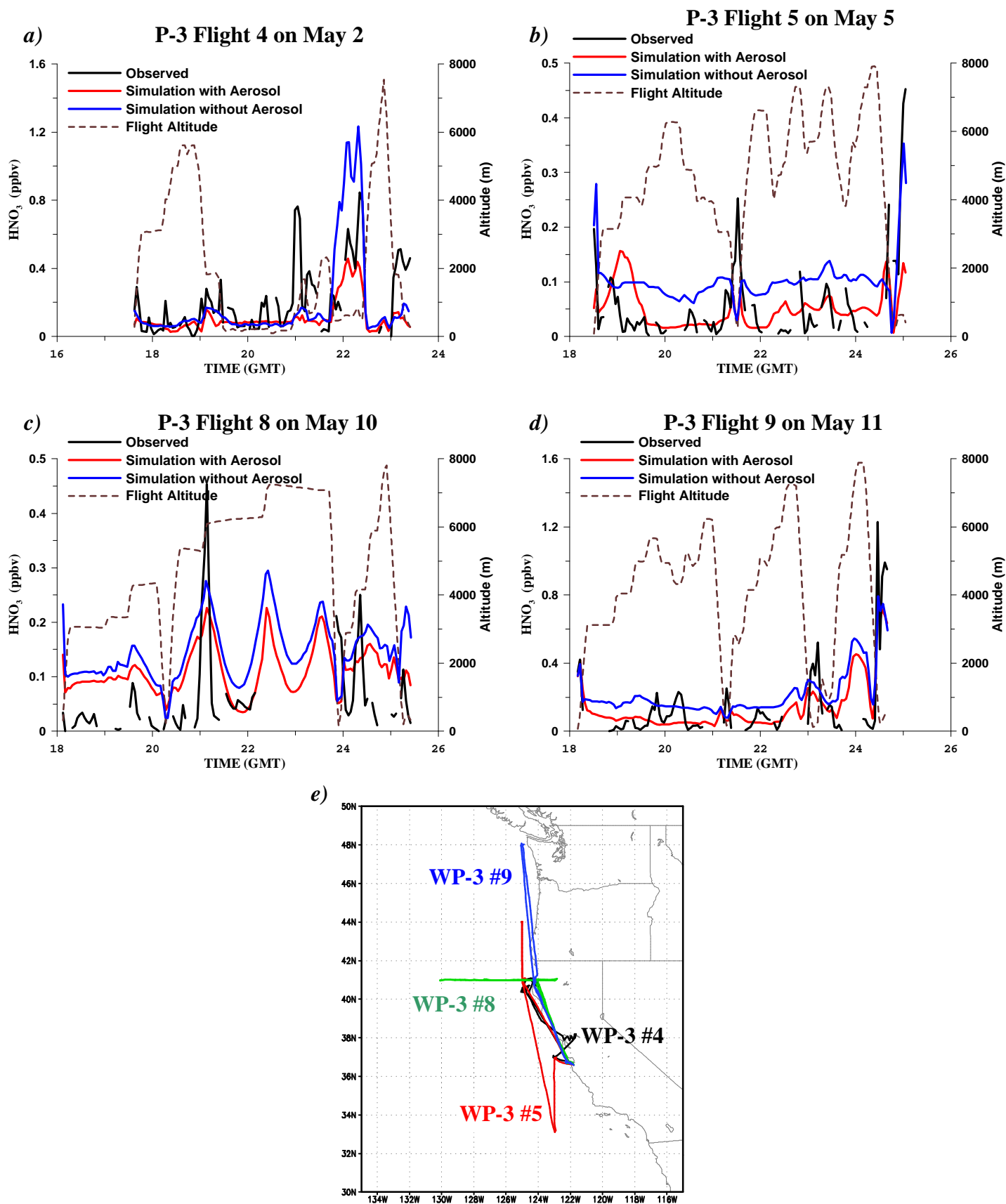
Figure 8

Figure 9

Figure 10

Figure 11

a) **Observed CO (ppbv)**

b) **Simulated CO (ppbv)**

c) **CO Difference (ppbv) (simulated - observed)**

d) **Simulated VOC Age (hour)**

e) **Observed NO$_y$ (ppbv)**

f) **NO$_y$ Difference (ppbv) (simulated - observed)**

Figure 12

Figure 13